

Nova Southeastern University NSUWorks

Biology Faculty Articles

Department of Biological Sciences

6-2023

Potential Impacts of Environmental Bacteria on the Microbiota of Loggerhead (Caretta caretta) and Green (Chelonia mydas) Sea Turtle Eggs and Their Hatching Success

Colleen McMaken Nova Southeastern University, cm2508@mynsu.nova.edu

Derek Burkholder Nova Southeastern University, dburkholder@nova.edu

Rosanna J. Milligan Dr. Nova Southeastern University, rboyle@nova.edu

Jose V. Lopez Nova Southeastern University, joslo@nova.edu

Follow this and additional works at: https://nsuworks.nova.edu/cnso_bio_facarticles

Part of the Biology Commons

NSUWorks Citation

McMaken, Colleen; Derek Burkholder; Rosanna J. Milligan Dr.; and Jose V. Lopez. 2023. "Potential Impacts of Environmental Bacteria on the Microbiota of Loggerhead (Caretta caretta) and Green (Chelonia mydas) Sea Turtle Eggs and Their Hatching Success." *Microbiologyopen* 12, (3): e1363. doi:10.1002/mbo3.1363.

This Article is brought to you for free and open access by the Department of Biological Sciences at NSUWorks. It has been accepted for inclusion in Biology Faculty Articles by an authorized administrator of NSUWorks. For more information, please contact nsuworks@nova.edu.

DOI: 10.1002/mbo3.1363

ORIGINAL ARTICLE

WILEY

Potential impacts of environmental bacteria on the microbiota of loggerhead (Caretta caretta) and green (Chelonia mydas) sea turtle eggs and their hatching success

Halmos College of Arts and Sciences, Nova Southeastern University, Florida, Fort Lauderdale, USA

Correspondence

Colleen M. McMaken, Halmos College of Arts and Sciences, Nova Southeastern University, Fort Lauderdale, FL, USA. Email: colmcmaken@gmail.com

Funding information

President's Faculty Research and Development, Nova Southeastern University, Grant/Award Number: #334882

Colleen M. McMaken 💿 | Derek A. Burkholder | Rosanna J. Milligan | Jose V. Lopez

Abstract

Sea turtle hatching success can be affected by many variables, including pathogenic microbes, but it is unclear which microbes are most impactful and how they are transmitted into the eggs. This study characterized and compared the bacterial communities from the (i) cloaca of nesting sea turtles (ii) sand within and surrounding the nests; and (iii) hatched and unhatched eggshells from loggerhead (Caretta caretta) and green (Chelonia mydas) turtles. High throughput sequencing of bacterial 16S ribosomal RNA gene V4 region amplicons was performed on samples collected from 27 total nests in Fort Lauderdale and Hillsboro beaches in southeast Florida, United States. Significant differences were identified between hatched and unhatched egg microbiota with the differences caused predominately by *Pseudomonas* spp., found in higher abundances in unhatched eggs (19.29% relative abundance) than hatched eggs (1.10% relative abundance). Microbiota similarities indicate that the nest sand environment, particularly nest distance from dunes, played a larger role than the nesting mother's cloaca in influencing hatched and unhatched egg microbiota. Pathogenic bacteria potentially derive from mixed-mode transmission or additional sources not included in this study as suggested by the high proportion (24%-48%) of unhatched egg microbiota derived from unknown sources. Nonetheless, the results suggest Pseudomonas as a candidate pathogen or opportunistic colonizer associated with sea turtle egg-hatching failure.

KEYWORDS

beach sand, eggs, hatching success, Pseudomonas, sea turtle

1 | INTRODUCTION

Sea turtles (Order Testudines) have been considered iconic marine reptiles due to their graceful beauty, navigational prowess, size, and longevity. Once numbering in the millions according to voyage logs from the 1400 to 1600s (Jackson, 1997), sea turtle populations have dwindled to the thousands, and six of seven extant sea turtle species have been listed as endangered or threatened globally. Sea turtles are highly migratory marine reptiles who nest on tropical beaches and are found globally in temperate and tropical waters (Balazs, 1980; Snell & Fritts, 1983; Spotila, 2004). The reproductive strategy among the different species of sea turtles is similar; all are oviparous with high

_____ This is an open access article under the terms of the Creative Commons Attribution License, which permits use, distribution and reproduction in any medium, provided the original work is properly cited.

© 2023 The Authors. MicrobiologyOpen published by John Wiley & Sons Ltd.

reproductive output to compensate for high natural mortality during early developmental stages (Butler, 1997; Wyneken et al., 1988). The eggs are fertilized internally, developing a soft shell before oviposition (Owens, 1980). After fertilization, females nest on sandy beaches depositing two to seven clutches of around 100 eggs per season (Carr, 1967). The eggs then incubate within the nest for 40–60 days, after which the hatchlings synchronize for a nighttime emergence (Carr & Hirth, 1961).

Survival rates of sea turtle eggs and hatchlings have decreased by different factors: physical nest destruction, predation, poaching, abiotic nest conditions (e.g., temperature, gas exchange, and moisture), and microbial interactions (Honarvar et al., 2011). Microbes live symbiotically with most eukaryotic organisms throughout their lifecycle (McFall-Ngai et al., 2013), and this includes sea turtles from egg formation up to and throughout adulthood. The primary mechanisms for microbial introduction are hypothesized to be through either (i) maternal transmission during the two-week formation period within the uterine tube and oviposition (Funkhouser & Bordenstein, 2013) or (ii) environmental transmission from sand surrounding the nest (Craven et al., 2007).

Maternal transmission is considered one of the main drivers of microbial introduction to sea turtle eggs. Sea turtle eggs spend the majority of their development (approximately two months) incubating in the sand without paternal care, therefore maternal transmission of bacteria can only occur during the two-week formation period before oviposition (Miller, 1985). Sea turtle hatchlings have been found to acquire a portion of their normal flora from the mother before oviposition (Scheelings, 2019). The cloaca can potentially also introduce infectious microbes due to its function as a combined opening for the digestive, urinary, and reproductive systems in birds. reptiles, and amphibians. The turtle cloaca has been shown to hold a highly complex and variable microbiome (Al-Bahry et al., 2009). Reproductive behaviors, environmental conditions, and diet can expose the cloaca to new and potentially pathogenic microbes. However, the establishment of new microbes in the cloaca depends on the presiding microbial community structure, the host's immune system, and the cloacal environment (Poiani, 2010). Additionally, mucus excreted during oviposition may contain antimicrobial properties that protect the embryos from potential pathogens in the environment or the reproductive tract (Keene, 2012; Praja et al., 2021; Soslau et al., 2011). This suggests that the transmission of microbes could occur at any time, whether it be before deposition (from the mother) or later on in the incubation process (from the environment).

Environmental transmission is considered another main driver of microbial introduction due to eggshell characteristics (Craven et al., 2007; De Reu et al., 2006). Sea turtles have a flexible eggshell composed of an inner organic membrane and an outer inorganic, calcareous layer that both contain numerous openings allowing for moisture and respiratory gas transfer (Al-Bahry et al., 2009; Chan & Solomon, 1989). Environmental conditions in nests (e.g., temperature and oxygen content) have been found to correlate with the microbial assemblage present in the nest and with overall hatching success

(Bézy et al., 2015). Physical properties of the nest substrate (e.g., sand grain size and organic matter content) establish diffusion rates of gases (e.g., oxygen and carbon dioxide), transport of water, and transmission of heat throughout the egg chamber, affecting developmental rates of the egg clutch (Ackerman, 1996; Mortimer, 1990). Biotic factors (e.g., clutch size and microbial activity) can alter nest temperature and gas composition, which can indirectly affect hatching success (Ackerman, 1974).

Beach sands harbor diverse microbiomes that provide important ecosystem services, such as water purification, biogeochemical cycling, and organic compound mineralization (Boehm et al., 2014). Geographic and tidal zone locations, sand depth, water pollution, atmospheric deposition, and anthropogenic activity have been found to cause sand microbial variation between beaches (Boehm et al., 2014; Hu et al., 2019; Piggot et al., 2012; Staley & Sadowsky, 2016). Sand microbes from within sea turtle nests can also differ from the beach sand microbes due to the mucus excreted by the female providing a comparatively warm, moist environment and potential nutrient resource that promotes the growth of certain microbes within the egg chamber (Honarvar et al., 2011; Wyneken et al., 1988). When hatched and unhatched (also referred to as "failed") egg microbiomes were previously characterized, differences in microbial species and overall microbial abundances were found to impact hatching success. Higher bacterial and fungal abundances and diversity in eggs (Gifari et al., 2018; Wyneken et al., 1988) and sand (Bézy et al., 2015; Honarvar et al., 2011) have been found to negatively impact the hatch rate success of a nest.

Research in sea turtle egg hatchability and their association with pathogens has been focused on sea turtle egg fusariosis (STEF) and Fusarium solani species complex (FSSC), which are fungal diseases causing high egg mortality in sea turtle nests worldwide (Gleason et al., 2020; Rosado-Rodríguez & Maldonado-Ramírez, 2016; Sarmiento-Ramírez et al., 2010, 2014, 2017; Sidigue et al., 2017; Smyth et al., 2019). STEF and FSSC are not the only fungi capable of penetrating and infecting sea turtle eggs. Additional fungal species identified in sea turtle eggs include Aspergillus, Emericella, Rhizopus, Actinomucor, and Apophysomyces (Candan, 2018). With research mostly focused on the fungal diseases impacting sea turtle hatching success, there is a lack of information on the role of bacteria within sea turtle eggs. Sarmiento-Ramírez et al. (2014) provided the first analysis of bacterial composition in sea turtle eggs using PhyloChip methods. However, their study had a limited sampling size of four eggs (two hatched and two unhatched) collected from two nests and was focused on Fusarium-infected nests. More research using new culture-independent methods (e.g., massive parallel sequencing) and larger sampling efforts need to be conducted on uninfected nests to help determine the potential role of bacteria in sea turtle egg development.

To date, only a few studies have utilized high throughput sequencing (HTS) of metagenomic DNA to characterize sea turtle egg microbiota and correlate it with hatching success. Scheelings (2019) used HTS to determine how sea turtle hatchlings acquire their normal flora by comparing bacterial composition between freshly laid eggs and blood and gut samples from nesting mothers and hatchlings. Their results found that the eggs share 44% of their microbiota with the mother's blood, suggesting that a portion of their microbiota is transferred before shell formation (Scheelings, 2019). Hoh et al. (2020) used HTS to compare hatchery practices by the differences in bacterial and fungal pathogens found in the sand and Fusariuminfected eggs. Their results found that eggs suffer an increased risk of infection in hatcheries that reuse sand for several nesting seasons (Hoh et al., 2020). Bézy et al. (2020) used HTS to compare the microbial composition in nest sand between areas of different embryo survivorship on a Costa Rica arribada beach but did not investigate the egg and cloaca microbiota. Their results found that sand microbial composition corresponds to particular environmental conditions suggesting that the presence of pathogenic microbes alone cannot fully determine hatching success (Bézy et al., 2020). Vecchioni et al. (2022) used HTS to characterize and compare loggerhead sea turtle egg microbiota along the Italian coasts of the Mediterranean Sea. Their results suggested that egg microbiota are shaped by maternal and environmental influences alongside a protective role of eggshells, but their study lacked microbial sampling from the nesting females to validate this claim (Vecchioni et al., 2022). Capri et al. (2023) used HTS and culture-dependent methods to determine the bacteria and fungi responsible for hatching failure in two green sea turtle nests with different hatching success rates (0% and 59%). Their results found that differences in bacterial abundance may have a more predominant role in hatching success than fungi. Additionally, their results found that Pseudomonas and Brucella were the main bacteria affecting hatching success. Capri et al. (2023) also hypothesized that Pseudomonas derived from the sand while Brucella from the nesting female, however, their study lacks bacterial analysis from the nesting female to support this claim.

The primary aim of our study was to characterize the bacterial composition of hatched and unhatched eggs for two important turtle species, the loggerhead (*Caretta caretta*) and green (*Chelonia mydas*) turtles in southeast Florida, and compare them with samples taken from the cloaca of the nesting sea turtles and from the sand surrounding and within the nests. Identifying microbial differences between hatched and unhatched eggs from loggerhead and green turtles will lead to additional insights into the potential roles of bacteria in sea turtle eggs (e.g., commensals or pathogens). Comparisons between cloaca, sand, and egg samples will also provide new information on the possible transmission source for sea turtle egg bacterial microbiota.

2 | EXPERIMENTAL PROCEDURES

2.1 | Data acquisition

Sampling occurred during the 2021 nesting season (May–October) with help from the Broward County Sea Turtle Conservation Program (BCSTCP). Sampling was authorized by the Florida Fish and Wildlife

Conservation Commission (FWC) under Marine Turtle Permit (MTP) #255. A total of 20 nests were sampled for loggerhead (*Caretta caretta*) and 7 nests for green (*Chelonia mydas*) turtles. The nests were determined when nesting turtles were identified by the BCSTCP staff under MTP #255 in Hillsboro and Fort Lauderdale beaches in Broward County, Florida. Broward County is located in southeast Florida and serves as a consistent annual nesting site for loggerheads (*C. caretta*), green turtles (*C. mydas*), and to a lesser extent leatherbacks (*Dermochelys coriacea*), which constitute approximately 90%–95%, 5%, and 1% of local nesting, respectively (Burkholder et al., 2020).

A total of nine different samples were collected from each nest (one cloaca, two sand, and six egg samples) resulting in 243 total samples being collected. Cloaca swabs were collected after oviposition while the turtle was temporarily detained by BCSTCP staff for research purposes (e.g., tagging and isotope and genetic sampling). A BD BBL[™] CultureSwab[™] EZ sterile media-free polyurethane foam swab was dipped into sterile water and inserted into the cloaca (about 5-6 cm or no further than the point of resistance) where the epithelium was gently scraped for about 10 s. The remaining samples were collected following the nest excavation by the BCSTCP staff under MTP #214, which occurred three days after the nest hatch-out had been documented following FWC Marine Turtle Guidelines. Two sand samples were collected from each nest using a sterile 2 mL microcentrifuge collection tube: the first ("nest sand") was collected from the compact sand found at the bottom of the egg chamber, while the second ("control sand") was collected from a human dug area of sand approximately 1-2 m away from the nest at the same depth as the egg chamber (range = 41-80 cm; Table 1). The latter sample was meant to represent sand samples with no association to any sea turtle nests (the original beach bacterial community). Within each nest, three hatched eggs and three unhatched eggs were swabbed on the interior portion of the eggshell for up to 30 s. Only hatched eggs that had more than 50% of the eggshell intact after hatchling emergence were selected since the interior could easily be distinguished from the exterior side. These hatched eggs also had a reduced chance of sand bacterial contamination affecting the eggshell's interior bacterial community due to the collapsing of the shell structure after turtle emergence. Only unhatched eggs that were not fully white, turgid, or demonstrated other characteristics suggestive of the presence of a living embryo were opened for swabbing. Eggs that were determined to be potentially viable were returned to the nest and reburied. All samples were placed on ice and transported to the laboratory to be stored at -80°C until DNA extractions began.

Environmental metadata were collected following standards set by Knight et al. (2012). At new nest sites, BCSTCP staff recorded the GPS location of the egg chamber, documented final nesting treatment (in situ [nest left where it was originally laid] or relocated [nest moved to a new location due to unsafe conditions]), and measured the nest's horizontal distance (m) from high tide and dunes using a Trimble GeoExplorer 6000 Series or Trimble GeoExplorer 2008 Series. Throughout the remainder

| Nest number | Turtle species | Beach | Latitude | Longitude | Bacterial incubation length (days) | Hatching success (%) | Clutch size | Chamber depth (cm) | High tide distance (m) | Dune distance (m) | Washover occurrences | pH (side, bottom) | Temperature (°C) (side, bottom) | Conductivity (µS/cm) (side, bottom) | Sand grain size (µm) (| Sand sorting Sofficient So) |
|----------------|----------------|-----------------|-----------|------------|---|----------------------------|-------------|-----------------------|------------------------------|-------------------------|-------------------------|----------------------|---------------------------------------|---|------------------------------|-----------------------------------|
| 127 | Loggerhead | Hillsboro | 26.266364 | -80.079940 | 54 | 89.15 | 129 | 73 | 1.8 | 4.6 | 0 | 8.16, 8.13 | 31.8, 31.6 | 2, 52 | 442.984 | 1.286 |
| 236 | Loggerhead | Hillsboro | 26.271728 | -80.079553 | 55 | 86.49 | 111 | 69 | 8.5 | 4.3 | 0 | 8.42, 7.22 | 33.9, 32.6 | 10, 66 | 399.827 | 1.269 |
| 242 | Loggerhead | Hillsboro | 26.301084 | -80.076934 | 54 | 50.00 | 98 | 48 | 4.3 | 9.1 | 0 | 8.41, 8.04 | 31.9, 32.0 | 8, 35 | 344.769 | 1.315 |
| 282 | Loggerhead | Hillsboro | 26.273437 | -80.079461 | 53 | 37.41 | 139 | 63 | 10.4 | 0.9 | 0 | 8.32, 7.57 | 33.5, 32.6 | 5, 20 | 465.131 | 1.445 |
| 322 | Loggerhead | Hillsboro | 26.281690 | -80.078690 | 56 | 86.76 | 136 | 65 | 7.3 | 7.6 | 0 | 8.41, 8.24 | 32.5, 32.5 | 5, 35 | 414.71 | 1.276 |
| 323 | Loggerhead | Hillsboro | 26.291539 | -80.077853 | 56 | 92.63 | 95 | 54 | 5.8 | 8.5 | 2 | 8.26, 8.36 | 32.2, 32.5 | 52, 21 | 368.333 | 1.282 |
| 392 | Loggerhead | Hillsboro | 26.282356 | -80.078671 | 54 | 91.13 | 124 | 53 | 11.6 | 3.4 | 0 | 9.04, 8.37 | 32.9, 33.1 | 2, 36 | 369.319 | 1.277 |
| 395 | Loggerhead | Hillsboro | 26.294768 | -80.077692 | 52 | 100.00 | 89 | 52 | 8.2 | 3.7 | 0 | 8.67, 8.41 | 33.5, 33.3 | 3, 36 | 349.142 | 1.296 |
| 456 | Loggerhead | Fort Lauderdale | 26.105561 | -80.104279 | 55 | 92.62 | 122 | 54 | 8.5 | 80.5 | 0 | 8.50, 8.83 | 32.7, 33.0 | 3, 53 | 565.06 | 1.744 |
| 457 | Loggerhead | Fort Lauderdale | 26.103821 | -80.104730 | 53 | 85.82 | 134 | 41 | 45.4 | 0.0 | 0 | 7.82, 8.45 | 32.4, 32.8 | 3, 131 | 450.679 | 1.556 |
| 463 | Loggerhead | Hillsboro | 26.288877 | -80.078165 | 54 | 93.75 | 112 | 75 | 14.0 | 1.2 | 0 | 7.87, 8.08 | 31.9, 31.8 | 1, 16 | 538.46 | 1.662 |
| 463 | Loggerhead | Fort Lauderdale | 26.096287 | -80.104877 | 55 | 95.58 | 113 | 63 | 12.2 | 110.0 | 1 | 8.16, 7.68 | 31.8, 32.2 | 91, 230 | 176.818 | 1.189 |
| 519 | Loggerhead | Fort Lauderdale | 26.099545 | -80.104470 | 55 | 79.17 | 85 | 54 | 0.0 | 162.2 | 0 | 6.78, 6.78 | 32.0, 32.3 | 6, 70 | 580.429 | L.844 |
| 520 | Loggerhead | Fort Lauderdale | 26.099477 | -80.104516 | 53 | 89.90 | 66 | 49 | 3.4 | 154.5 | 0 | 6.78, 6.76 | 30.1, 31.3 | 7, 46 | 459.551 | L.779 |
| 521 | Loggerhead | Fort Lauderdale | 26.097303 | -80.104628 | 57 | 54.26 | 94 | 54 | 7.3 | 186.8 | 23 | 6.70, 6.73 | 29.7, 32.4 | 242, 2000 | 352.39 | 1.502 |
| 522 | Loggerhead | Fort Lauderdale | 26.095800 | -80.104921 | 52 | 92.86 | 126 | 60 | 5.8 | 87.8 | 2 | 6.77, 6.82 | 29.7, 31.0 | 208, 133 | 473.56 | 1.836 |
| 697 | Loggerhead | Fort Lauderdale | 26.264892 | -80.080085 | 50 | 95.05 | 101 | 51 | 4.6 | 95.4 | т | 6.79, 6.78 | 30.5, 30.8 | 351, 140 | 431.005 | l.541 |
| 698 | Loggerhead | Fort Lauderdale | 26.266499 | -80.079955 | 53 | 83.00 | 100 | 53 | 10.7 | 112.8 | 0 | 7.81, 7.83 | 32.4, 33.1 | 4, 48 | 394.668 | 1.382 |
| 669 | Loggerhead | Fort Lauderdale | 26.104836 | -80.104836 | 51 | 79.07 | 129 | 48 | 15.8 | 159.7 | 0 | 8.17, 8.04 | 32.5, 32.7 | 6, 16 | 421.679 | 1.686 |
| 892 | Loggerhead | Fort Lauderdale | 26.102794 | -80.104490 | 52 | 85.38 | 130 | 60 | 9.8 | 9.8 | 0 | 7.86, 7.85 | 29.5, 30.0 | 12, 70 | 425.691 | 1.248 |
| 695 | Green | Hillsboro | 26.099619 | -80.104584 | 56 | 96.03 | 126 | 67 | 10.4 | 0.0 | 0 | 7.01, 6.92 | 30.5, 31.6 | 19, 236 | 377.553 | 1.273 |
| 696 | Green | Hillsboro | 26.168980 | -80.097780 | 53 | 95.24 | 84 | 80 | 13.4 | 0.0 | 0 | 6.67, 6.77 | 33.3, 32.5 | 6, 59 | 477.532 | l.551 |
| 823 | Green | Fort Lauderdale | 26.171648 | -80.097618 | 52 | 92.56 | 121 | 68 | 13.4 | 6.4 | 0 | 6.41, 6.76 | 31.6, 31.4 | 3, 107 | 496.768 | L.772 |
| 867 | Green | Fort Lauderdale | 26.153350 | -80.100218 | 55 | 95.61 | 114 | 74 | 39.0 | 0.0 | 0 | 7.94, 7.93 | 28.6, 29.3 | 2, 113 | 457.801 | L.549 |
| 1166 | Green | Hillsboro | 26.264360 | -80.080120 | 57 | 97.43 | 116 | 77 | 11.6 | 0.0 | 0 | 7.6, 7.39 | 28.5, 29.2 | 2, 24 | 435.663 | 1.283 |

4 of 34

TABLE 1 Metadata associated with nests sampled.

| nd sorting sfficient | 8 | 90 |
|---|------------|---------------|
| Sar coe tm) (S ₀) | 51 1.3 | 35 1.3 |
| Sand grain size (µ | 409.3 | 341.5 |
| Conductivity (μS/cm) (side, bottom) | 4, 67 | 16, 11 |
| Temperature (°C) (side, bottom) | 31.0, 32.5 | 28.6, 29.5 |
| pH (side, bottom) | 8.06, 8.12 | 8.08, 7.75 |
| Washover occurrences | 0 | 0 |
| Dune distance (m) | 0.0 | 0.0 |
| High tide distance (m) | 18.9 | 11.3 |
| Chamber depth (cm) | 71 | 60 |
| Clutch size | 121 | 116 |
| Hatching success (%) | 95.04 | 95.69 |
| Bacterial incubation length (days) | 50 | 55 |
| Longitude | -80.079680 | -80.079620 |
| Latitude | 26.269120 | 26.271610 |
| Beach | Hillsboro | Hillsboro |
| Turtle species | Green | Green |
| Nest number | 1167 | 1168 |
| | | |

to the topmost point of leveled sand. High tide and dune distance (m) were measured by BCSTCP staff when the nest was originally laid. Washover occurrences (50% of the nest was exposed to tidal action) were observed by BCSTCP staff throughout each nest's incubation Note: Nest number indicates the nest number associated with the nest sampled on each beach, as determined by the BCSTCP. Turtle species (Loggerhead or Green Turtle), beach (Fort Lauderdale or Hillsboro), and GPS location (latitude and longitude) associated with each nest number are identified. Bacterial incubation length (days) was determined by the number of days between the date the nests were originally (total number of hatched, LPIP, and pH were collected from the side and bottoms of each nest (both values shown in the table) immediately and DPIP eggs) by the total clutch size egg chamber. chamber 75 g of sand collected from within each aid and the date of excavation. Hatching success was determined by dividing the number of hatched eggs (total number of hatched, LPIP, egg measuring from the bottom of the Conductivity, temperature, coefficients were determined using were removed, depth was measured after all nest contents period, total amounts of occurrences were calculated after nest removal. Sand grain size and sorting Broward County after nest contents were removed. Chamber Abbreviations: BCSTCP, DPIP, and whole eggs).

live pipped hatchlings dead pipped hatchlings; LPIP, Sea Turtle Conservation Program; DPIP, MicrobiologyOpen

of the incubation period, nest inundation and any instances of predation were recorded. Three days after the initial nest hatchout was marked by BCSTCP staff, the nest was excavated to determine hatching success through egg counts. The egg counts included the number of hatched eggs, live hatchlings in nest (LIN), dead hatchlings in nest (DIN), live pipped hatchlings (LPIP), dead pipped hatchlings (DPIP), and whole unhatched eggs. These were used to determine hatching success by dividing the number of hatched eggs (total number of hatched, LPIP, and DPIP eggs) by the total clutch size (total number of hatched, LPIP, DPIP, and whole eggs).

When all eggs were removed from the egg chamber, egg chamber depth (cm) was recorded. Nest sand pH (Hanna Instruments[®] HI981030 GroLine Soil pH Tester) and temperature and conductivity (Hanna Instruments[®] HI98331 Soil Test[™]) were measured at the bottom of the egg chamber and along the middle of the egg chamber wall to determine the differences experienced throughout the egg chamber. Sand grain size and sorting coefficient were determined by collecting 75 g of nest sand from each nest location into sterile 50 mL FalconTM centrifuge tubes. Sand samples were fractionated with a set of sieves (63, 125, 250, 500, 2000, and 4000 microns) to determine the particle-size distribution by mass. The "G2Sd" R package (Fournier et al., 2014) was used to calculate the mean of the grain-size distribution (geometric method of moments; Bunte & Abt, 2001) and sorting coefficient using the Trask Index (Trask, 1932) defined as D25/D75 (mm scale).

2.2 **Bacterial sequencing**

16S rRNA gene amplicons were sequenced using Earth Microbiome Project protocols for the Illumina MiSeg platform (Thompson et al., 2017). Microbial DNA was extracted from the cloaca, sand, and egg samples following protocols for the QIAGEN DNEasy PowerLyzer PowerSoil Kit. After extraction, each sample was checked using electrophoresis on a 1% agarose gel to ensure that the DNA was successfully extracted. Polymerase chain reaction (PCR) tests were run on the DNA isolates using Invitrogen[™] Platinum[™] Hot Start PCR Master Mix (2x). 515F and 806R primers were used to amplify the V4 region of the 16S rRNA gene (Caporaso et al., 2011; Easson & Lopez, 2019). The barcoded PCR products were then cleaned using Beckman Coulter AMPure XP beads, purifying the DNA amplicons away from contaminants such as dNTPs, salts, primers, and primer dimers (Greenwald et al., 2019). The final DNA concentration was checked using a Qubit[®] 2.0 Fluorometer and then diluted to a normalization of 4-5 ng/µL. Ten samples were then pooled at a time in equal volumes for a final quality check using Agilent 2200 and 4150 TapeStation Systems. A final library pool containing equal volumes of DNA products from all samples being sequenced was created and loaded into the Illumina MiSeq using the MiSeq Reagent Kit v3 at 600 cycles of sequencing following a modified Illumina workflow protocol used by the Microbiology and Genetics Laboratory at Nova Southeastern University Halmos College of Arts and Sciences.

2.3 | Data analysis

The Quantitative Insights into Microbial Ecology v.2 (QIIME 2 v.2021.8) pipeline was used to perform microbial bioinformatics after sequencing was completed (Bolyen et al., 2019). Raw sequences were demultiplexed and quality-filtered using DADA2 (Callahan et al., 2016). Amplicon sequence variants (ASV) were aligned with MAFFT (Katoh et al., 2002) and used to construct phylogeny with fasttree2 (Price et al., 2010). Taxonomy was assigned to the ASVs using the SILVA 138.1 feature classifier for the 515F/806R region of sequences (Quast et al., 2012; Yilmaz et al., 2014) after it was trained using scikit-learn 0.24.1 (Bokulich et al., 2018). Before analysis, the feature table created by QIIME 2 was cleaned using the "decontam" (Davis et al., 2018) and "phyloseq" packages (McMurdie & Holmes, 2013) in R (v.4.0.2) (R Core Team, 2020). Two DNA extraction controls were sequenced and used as a reference for contamination cleaning using the prevalence method with a threshold of 0.1, which removed 21 ASVs. Further cleaning was done without the incorporation of DNA extraction controls using the frequency method with a threshold of 0.1, which removed an additional 107 ASVs. A final cleaning check was performed using the "vegan" package (Oksanen et al., 2020) where ASVs occurring less than 0.001% were removed, resulting in two additional ASVs being removed. Low threshold and abundance levels were used to preserve the bacterial diversity in the samples.

Alpha diversity, including Margalef's species richness (Margalef, 1958), Shannon diversity (Shannon & Weaver, 1949), and Inverse Simpson's diversity (Simpson, 1949), was assessed for each sample in the dataset. Generalized linear mixed-effect models (GLMM) with gaussian distribution were performed using the "Ime4" R package (Bates et al., 2015) for all samples to determine which variables affected each alpha diversity metric (response variables). Before setting up the model, the variables were assessed for covariation. Variables with strong correlations included latitudelongitude (corr = 0.996), latitude—dune distance (corr = -0.791), longitude-dune distance (corr = -0.757), temperature at bottom of nest-temperature at side of nest (corr = 0.873), and pH at bottom of nest-pH at side of nest (corr = 0.868). From these correlations, latitude, longitude, temperature at side of nest, and pH at side of nest were removed from the model while dune distance, temperature at bottom of nest, and pH at bottom of nest were retained. The fixed variables used within the "full" model were turtle species (categorical with two levels), beach (categorical with two levels), sample type (categorical with five levels), incubation length (continuous), clutch size (continuous), hatching success (continuous), chamber depth (continuous), high tide distance (continuous), dune distance (continuous), temperature at bottom of nest (continuous), pH at bottom of nest (continuous), conductivity at side of nest (continuous), conductivity at bottom of nest (continuous), sand grain size (continuous), and sorting coefficient (continuous). The interaction terms were sample type-turtle species and sample type-beach. Each model also included the nest number as a random effect to control for lack of independence between eggs from the same nest. Final "best-fit"

models were determined through term-selection by AIC score comparisons. If the removal of a variable caused the AIC score to increase by a value of two or more than that variable was kept for the "best-fit" model. The "best-fit" models were validated by plotting Pearson residuals against fitted values, against each variable (covariate) in the model, and against each variable (covariate) not in the model. If the interaction terms were kept in the "best-fit" model, then the "emmeans" package (Lenth, 2023) was used to perform pairwise analyses to determine which levels of the categorical variables had a significant effect.

To assess if sampling effort affected the differences in alpha diversity metrics between turtle species and beaches, sample-sizebased rarefaction (interpolation) and prediction (extrapolation) curves were created using the "iNEXT" R package (Hsieh et al., 2016) by computing diversity estimates using the number of samples with respect to the total number of individuals (ASVs). These analyses (and future beta diversity analyses) were completed for each turtle species and beach separately. Comparisons between beaches were only made using loggerhead turtle samples, due to the low number of green turtle nests sampled at Fort Lauderdale (n = 2).

The data were then converted into relative abundance (resulting in proportional data for each sample) to perform beta diversity analyses between turtle species and beaches. Bray-Curtis dissimilarity matrices with ASVs standardized by sample total were used to generate nonmetric multidimensional scaling plots using PRIMER v.7.0.17 software (Clarke & Gorley, 2015). One-way, unordered analysis of similarity (ANOSIM: Clarke, 1993) with 9999 permutations was used to determine whether there were differences in community composition between sample types (hatched eggs, unhatched eggs, nest sand, control sand, and cloaca), beaches, and turtle species. Where significant differences were detected between groups, Similarity Percentages (SIMPER) analyses (Rees et al., 2004) with 9999 permutations were then used to determine which bacterial taxa were significantly different between sample types, beaches, and turtle species. Hatched and unhatched egg microbiota were compared to sand and cloaca microbiota to determine potential transmission sources using the "venn" R package (Dusa, 2021) to determine how many unique ASVs hatched and unhatched eggs had in common with cloaca and sand samples. SourceTracker (Knights et al., 2011) was then used to estimate the proportional contribution of each source type for hatched and unhatched egg microbiota using Bayesian modeling for proposed known (cloaca, control sand, and nest sand samples) and unknown source environments. An additional SourceTracker analysis was used to estimate the proportional contribution of control sand and cloaca microbiota on nest sand microbiota.

Differences between sample types were further visualized in R with the "phyloseq" and "microbiome" (Lahti & Shetty, 2019) packages using the relative abundances of the most abundant phyla and genera. Canonical analysis of principal coordinates (CAP; Anderson & Willis, 2003) was conducted in PRIMER v7 with PERMANOVA+ (Anderson et al., 2008) using Bray-Curtis dissimilarity matrices analyzed against egg sample types. The CAP results were then overlayed with Pearson correlations to determine which bacteria best categorized and distinguished the hatched egg sample types from the unhatched egg sample types. To test if beach locale (GPS location) affected inter-nest bacterial variance, Mantel tests were performed on control sand and nest sand microbiota (Bray-Curtis distance matrices) with a matrix composed of straight-line distances between nests. BIO-ENV analyses (Clarke & Ainsworth, 1993) were conducted in "vegan" to correlate the relative abundance data (Bray-Curtis dissimilarity with ASVs standardized by sample total) with the environmental data (Euclidean distances with scaled data) to determine what environmental factors best explained patterns of bacterial community change. Mantel tests (Mantel & Valand, 1970) with Spearman's rank correlations and 9999 permutations were performed after BIO-ENV analyses to estimate significance.

3 | RESULTS

3.1 | Illumina MiSeq run and taxonomy statistics

Two Illumina MiSeq runs were used to sequence all 243 samples collected. The Q30 score for both runs was above 70% (77.21% and 70.42%) with more than 78% (85.63% and 78.9%) of clusters passing through Illumina's internal quality filtering procedure. On average, 117,698 (range: 46,217-237,925) high-quality 16S rRNA gene sequences (V4 regions) were obtained from the 243 samples collected (Table A1). A total of 16,516 taxa were identified after data cleaning (the original raw count was 16,646 taxa). An average of 251 unique ASVs (range: 28-1644 ASVs) were identified per sample (Table A1).

3.2 | Bacterial community compositions

Control sand samples had significantly greater bacterial taxa richness and diversity when compared to egg and cloaca sample types in loggerhead and Hillsboro samples (Figure 1 and Table A2). Nest sand samples were significantly greater in bacterial taxa richness and diversity for both turtle species and beaches only when compared to hatched and unhatched eggs (Figure 1 and Table A2). Between egg sample types, hatched eggs had greater bacterial taxa richness and diversity than unhatched eggs (Figure 1 and Table A2). Loggerhead cloaca and control sand samples were significantly greater in bacterial richness and diversity than green turtles (Figure 1 and Table A3), while green turtle unhatched egg samples were significantly greater in diversity than loggerheads (Figure 1 and Table A3). Hillsboro control sand samples were significantly greater in bacterial richness and diversity than Fort Lauderdale (Figure 1 and Table A3). Although more samples were collected from loggerhead nests, these trends remained consistent when sample-size was taken into consideration (Figures A1 and A2). One-way ANOSIM tests showed that there were significant differences between the sample types in separate analyses

for each turtle species (Loggerhead: R = 0.493, p < 0.001; Green Turtle: R = 0.154, p = 0.006) and each beach (Hillsboro: R = 0.659, p < 0.001; Fort Lauderdale: R = 0.339, p < 0.001). Nest sand microbiota were found to be more similar to hatched and unhatched egg microbiota than control sand and cloaca microbiota in both turtle species and beach comparisons (Figure 2 and Table A4). Interestingly, nest sand and control sand samples were found to be significantly different from one another in both turtle species and beach analyses (Figure 2 and Table A4).

3.3 | Differences between hatched and unhatched egg microbiota

One-way ANOSIM tests found significant differences in bacterial taxa between hatched and unhatched eggs in both turtle species (Loggerhead: R = 0.345, p < 0.001; Green: R = 0.132, p < 0.001) and beaches (Hillsboro: R = 0.421, p < 0.001; Fort Lauderdale: R = 0.262, p < 0.001). Loggerheads had a greater dissimilarity between hatched and unhatched egg microbiota than green turtles at the phylum (31.52% and 18.17%, respectively) and genus-levels (77.00% and 43.68%, respectively) (Tables A5 and A6). Loggerhead hatched and unhatched egg microbiota were found to be more dissimilar within Hillsboro than Fort Lauderdale beaches at the phylum (34.20% and 29.01%, respectively) and genus-levels (80.86% and 72.39%, respectively) (Tables A5 and A6).

Pseudomonadota (formerly known as Proteobacteria [Oren & Garrity, 2021) was found as the highest correlated phylum to the differences between hatched and unhatched eggs for both turtle species and beaches and was found in higher abundances in unhatched eggs except in green turtle analyses (Figure 3 and Table A5). Actinomycetota (formerly known as Actinobacteria [Oren & Garrity, 2021]) and Bacillota (formerly known as Firmicutes [Oren & Garrity, 2021) contributed higher abundances in unhatched eggs than hatched eggs (except for Bacillota in Hillsboro samples) while Bacteroidota contributed a higher abundance in hatched eggs than unhatched eggs (Figure 3 and Table A5). SIMPER analyses found Pseudomonas as the highest correlated genus to the differences between unhatched and hatched egg microbiota for both turtle species and beaches and was found in higher abundances in unhatched eggs (Figure 4 and Table A6). Alcaligenes also contributed a higher abundance in unhatched eggs than hatched eggs while Nitratireductor, Flavobacterium, Paenibacillus, and Sphingobacterium contributed higher abundances in hatched eggs (Figure 4 and Table A6).

Since hatched and unhatched eggs were found to be significantly different, canonical analysis of principal coordinates (CAP) was used to characterize the bacterial differences between the egg sample types. The CAP routine successfully separated egg samples with 93.827% (m = 39) correct classification. Only 1 out of 84 hatched eggs was misclassified, while 9 out of 78 unhatched eggs were misclassified. These CAP diagnostics indicate that the two sample types are distinct enough from one another to be used for predictive

WILFY_MicrobiologyOpen



FIGURE 1 Alpha diversity comparison based on sample type. Three alpha diversity metrics were visualized: Margalef's species richness (top), Shannon diversity (middle), and Inverse Simpson's diversity (bottom). Facets were created by sample type (cloaca, control sand, hatched egg, nest sand, and unhatched egg) with each box within the facets representing a different dataset (loggerhead [dark grey], green turtle [grey], Fort Lauderdale [light grey], and Hillsboro [white]). The box portion shows the interquartile range separated by the median (black horizontal line) with whiskers (black vertical lines) showing the variability outside the upper and lower quartiles and outliers represented as black dots. Statistically significant differences, represented by the asterisks (*) were determined using the pairwise test in the "emmeans" package after the "best-fit" model was determined (*p* values can additionally be found in Table A3). Cloaca samples were significantly different in bacterial richness and diversity between turtle species and were only significantly different between beaches (loggerhead samples only) for bacterial richness. Control sand samples were significantly different in bacterial richness and diversity between turtle species and beaches, while nest sand samples were only significantly different in bacterial richness between beaches. Unhatched eggs were significantly different in diversity between turtle species, while hatched eggs exhibited no significant differences.

modeling and characterizing differences between egg-hatching success. At the phylum level, hatched eggs were correlated (corr >0.3) with Bacteroidota, Verrucomicrobiota, Myxococcota, and Bdellovibrionota, while unhatched eggs were correlated with

Pseudomonadota and Bacillota (Table A7). At the genus level, hatched eggs were correlated (corr >0.55) with *Sphingopyxis*, *Pseudonocardia*, *Devosia*, and *Cohnella*, while unhatched eggs were correlated with *Pseudomonas* (Table A7).



FIGURE 2 Bray-Curtis dissimilarity nMDS comparisons of sample types in Loggerhead (a), Green Turtle (b), Fort Lauderdale (c), and Hillsboro (d) Samples. The similarity between sample types (cloaca [grey], control sand [orange], hatched egg [red], nest sand [gold], and unhatched egg [blue]) was visualized by nMDS plots in a two-dimensional frame, using a Bray-Curtis resemblance matrix to rank order the dissimilarity among pairs of samples. Stress values indicate the extent of the distortion to visualize the results in a two-dimensional frame. All stress values were below 0.2, indicating a good representation of the data. Within the plots, the closer points (samples) are to each other, the more similar the community composition. In all plots, hatched eggs were closer to the nest sand samples, while the unhatched egg samples were closer to some of the cloaca samples instead of sand samples. Cloaca and control sand samples were separated further away from nest sand and egg samples. nMDS, nonmetric multidimensional scaling.

3.4 Potential transmission source

Cloaca, sand, and egg samples were compared to determine whether there was a greater maternal or environmental effect on egg microbiota. All egg samples shared more unique ASVs with nest sand samples (3%-20%) than the cloaca (0.5%-2%) or control sand (1%-2.6%) samples (Figure 5). Unhatched eggs shared more unique ASVs than hatched eggs with the cloaca samples (+0.21%-1.12%) while hatched eggs shared more with the control (+0.22%-0.46%) and nest (+4.05%-16.91%) sand samples (Figure 5). In green turtle samples, however, unhatched eggs shared more unique ASVs with all other sample types than hatched eggs (Figure 5). For both turtle species and beach comparisons, there were still 24%-37% unique ASVs within hatched and unhatched egg samples that did not match any ASVs found in the other sample types (Figure 5), indicating there may be another potential source for bacterial transmission. Most sediment habitats display some of the highest bacterial diversities, which is partially reflected by lower than expected sharing of ASVs between control sand and nest sand in loggerhead samples (Figure 2 and Table A4).

SourceTracker analyses showed that nest sand most likely accounted for 61%-90% of bacterial introduction in hatched egg samples, identifying it as potentially the predominant source of bacterial communities (Figure 6). In unhatched egg samples, nest sand bacterial introduction only accounted for 44%-49% (Figure 6). A higher percentage of bacterial acquisition from an unknown source was identified in unhatched eggs (24%-48%) than hatched eggs (1%-4%; Figure 6). The cloaca was identified as the least likely origin of bacterial communities in sea turtle eggs, accounting for less than 1% of bacterial introduction in both egg sample types for all comparisons (Figure 6). The control sand was found to have a greater influence on green turtle egg microbiota (28%-36%) than loggerhead and beach comparisons (7%-13%; Figure 6). Since nest sand was identified as the predominant source of bacterial communities in egg samples, an additional SourceTracker analysis was run to determine if the nest sand had a stronger maternal (cloaca) or environmental (control sand) influence. Cloaca samples were found to account for only 0.89% of bacterial introduction in nest sand samples while control sand accounted for 46.55%, still, 52.56% of the nest sand bacterial source was identified from unknown sources.



FIGURE 3 Most abundant bacterial phyla bar chart comparison of egg samples for turtle species and beaches. The most abundant bacterial phyla (based on relative abundance) were identified separately for each turtle species (Loggerhead [a] and Green Turtle [b]) and for the loggerhead samples at each beach (Fort Lauderdale [c] and Hillsboro [d]). Only bacterial phyla that had a relative abundance greater than 1% in either hatched or unhatched egg samples were selected for representation in these bar charts. All egg samples contained about 99% of the most abundant phyla identified. Loggerhead and both beach samples had a higher abundance of Pseudomonadota in unhatched egg samples. Both turtle species and Fort Lauderdale samples had a higher abundance of Bacillota in unhatched egg samples. Bacteroidota and Actinomycetota were in higher abundance in hatched eggs than unhatched eggs for both turtle species and beaches. Bdellovibrionota was only identified as an abundant bacterial phylum in Hillsboro samples and had a higher relative abundance in hatched eggs (1.01%) than unhatched eggs (0.04%).

Additional analyses were run to further corroborate that sand sources may influence bacterial transmission into sea turtle eggs more than cloaca microbiota. Although cloaca samples were found to be significantly different between turtle species (one-way ANOSIM R = 0.561, p = 0.001), no significant differences in microbiota composition were found between turtle species for hatched or unhatched egg samples (one-way ANOSIM; unhatched egg: R = -0.069, p = 0.886; hatched egg: R = -0.118, p = 0.979). On the contrary, control sand, nest sand, hatched egg, and unhatched egg microbiota showed to be significantly different between beaches (one-way ANOSIM; nest sand: R = 0.158, p = 0.029; control sand: R = 0.274, p = 0.012; unhatched egg: R = 0.067, p = 0.032; hatched egg: R = 0.088, p = 0.002). A SIMPER analysis of the control sand showed that the main drivers of dissimilarity between beach microbiota were the genera Achromobacter, Alcaligenes, Brevibacillus, Nitratireductor, Ochrobactrum, Sphingobacterium, and Stenotrophomonas (Table A8).

3.5 | Environmental factors shaping egg microbiota

Geographic location was first analyzed to determine if it played a stronger role in shaping sea turtle egg microbiota than the beach environment at each nest. Although there was a significant correlation between nest distances and control sand bacterial variance (Mantel R = 0.247, p = 0.018), no significant correlation was found between nest distances and nest sand bacterial variance (Mantel R = 0.071, p = 0.158). This suggests that although beach sand bacterial communities change with geographic location, more environmental factors play a role in shaping the different bacterial communities within the nest. Therefore, the different variations in environmental factors present at each nest may have stronger effects on shaping egg microbiota.

Three environmental variables were retained with all interactive variables in the GLMM analysis of Margalef's species richness, which strongly explained the variability in bacterial species richness ($R^2 = 0.76$; Table A9). Incubation length, high tide distance, and temperature at the bottom of nests were all found to positively affect bacterial species richness (t = 1.01, 3.13, and 1.48, respectively). Three environmental variables were retained with all interactive variables in the GLMM analysis of Shannon diversity, which strongly explained the variability in bacterial diversity ($R^2 = 0.64$; Table A9). Clutch size, high tide distance, and temperature at the bottom of nests were all found to positively affect bacterial diversity (t = 3.65, 2.01, and 1.66, respectively). One environmental variable was

Wh fy



FIGURE 4 Most abundant bacterial genera bar chart comparison of egg samples for turtle species and beaches. The most abundant bacterial genera (based on relative abundance) were identified separately for each turtle species (Loggerhead [a] and Green Turtle [b]) and for the loggerhead samples at each beach (Fort Lauderdale [c] and Hillsboro [d]). Only bacterial genera that had a relative abundance greater than 1% in either hatched or unhatched egg samples were selected for representation in these bar charts. Loggerhead egg samples contained about 35% of the most abundant genera identified while green turtles were only composed of about 14%. Fort Lauderdale contained about 20% of the most abundant genera identified while Hillsboro was only composed of about 15%. Both turtle species and beaches had a higher abundance of *Pseudomonas* in unhatched egg samples while hatched eggs had higher abundances of *Sphingobacterium, Paenibacillus, Nitratireductor*, and *Lysobacter*.

retained with only the sample type—beach interactive variable in the GLMM analysis of Inverse Simpson's diversity, which moderately explained the variability in bacterial diversity ($R^2 = 0.25$; Table A9). Sorting coefficient was found to negatively affect bacterial diversity (t = -0.24).

BIO-ENV analyses showed that loggerhead egg samples had more environmental variables (six) explaining the change in bacterial communities when compared to green turtles (four) (Table A10). When analyzing beaches for loggerhead samples only, BIO-ENV analyses showed that Hillsboro samples had more environmental variables (six) explaining the change in bacterial communities when compared to Fort Lauderdale (two) (Table A10). BIO-ENV analyses determined that sand grain size for loggerhead and Fort Lauderdale samples (corr = 0.152 and 0.286, respectively), incubation length for green turtles (corr = 0.167), and conductivity at the side of nests (corr = 0.209) for Hillsboro samples were the single environmental variables which best explained correlation to bacterial community data.

Since significant differences were identified between hatched and unhatched eggs between beaches (and not between turtle species), a BIO-ENV analysis was performed on the control sand samples to determine what environmental factors shaped the beach microbiota. Dune distance was identified as the best environmental factor to explain the bacterial differences between Fort Lauderdale and Hillsboro (corr = 0.252). Nests in Fort Lauderdale were found to occur significantly (p < 0.001) further from the dunes than nests in Hillsboro (average distance from dune = 89.7 m and 3.1 m, respectively, Table 1).

4 | DISCUSSION

Our study presents the first in-depth analysis of sea turtle egg bacterial communities in the continental United States using molecular methods and indicates that the nesting environment plays a critical role as a source of egg bacteria during the incubation period in loggerhead and green turtles in this region. The comparative approach with molecular methods has revealed several striking features of sea turtle egg and nest microbiota. First, hatched and unhatched eggs have statistically different bacterial assemblages from each other, with unhatched eggs having a more variable microbiota consistent with dysbiotic environments. Unhatched eggs represent the ultimate dysbiosis for a microsymbiont. Second, the nest sand microbiota have a greater influence on egg microbiota than cloaca microbiota. We hypothesize that this may be due to the waning influence of the mother's cloaca on the egg microbiota and/or the greater exposure to the nest sand during the two-month incubation period before the eggs were sampled. Third, beach environments were found to significantly alter the egg microbiota, while turtle species had no statistically significant impact. Fourth, nest sand microbiota were distinct from control sand microbiota,



FIGURE 5 Unique ASV counts shared between sample types for Loggerhead (a), Green Turtle (b), Fort Lauderdale (c), and Hillsboro (d) samples. The number of shared ASVs between sample types (cloaca [black], control sand [orange], hatched egg [red], nest sand [gold], and unhatched egg [blue]) were identified separately for each turtle species (Loggerhead [a] and Green Turtle [b]) and for the loggerhead samples at each beach (Fort Lauderdale [c] and Hillsboro [d]). Unique ASVs hatched eggs shared with cloaca, control sand, and nest sand samples are highlighted in red, while unique ASVs unhatched eggs shared with cloaca, control sand, and nest sand samples are highlighted values were divided by the total number of ASVs (in parentheses under sample type name) to determine the percentage of the total composition of the unique ASVs made up of hatched and unhatched egg samples. In each Venn diagram, both hatched and unhatched eggs shared more unique ASVs with nest sand samples than either the cloaca or control sand. ASV, amplicon sequence variant.

potentially due to the introduction of the nesting female's mucus during oviposition.

4.1 | Hatched and unhatched egg microbiota comparisons

Our study was able to collect samples from nests with hatching success rates ranging from 37% to 100% (Table 1) and detected significant microbiota composition differences between hatched and unhatched eggs. Hatched eggs were found to have less variance in their microbiota when compared to unhatched eggs (Figure 2). Less

variance in hatched eggs may follow the so-called "Anna Karenina principle" (AKP). The AKP posits that healthy animals possess relatively stable, similar microbiomes, while this stability can be disrupted by a variety of external stressors in unhealthy animals, resulting in more variable microbiomes (Zaneveld et al., 2017).

4.2 | Hatched egg microbiota—The putative "healthy microbiota"

Based on the findings of the present study, hatched eggs correlated with having a lower species richness (between 300 and 400 ASVs)



FIGURE 6 Source bacterial contributions for hatched and unhatched eggs. Proportional contributions of known (cloaca [black], control sand [blue], and nest sand [gold]) and unknown (orange) bacterial source environments were estimated in hatched and unhatched eggs for loggerheads, green turtles, Fort Lauderdale (loggerhead samples only), and Hillsboro (loggerhead samples only). The majority of hatched egg microbiota were derived from the nest sand. Unhatched egg microbiota had a greater derivation from unknown sources than hatched eggs.

compared to unhatched eggs (>500 ASVs). At the phylum level, hatched (presumed healthy) eggs contained higher abundances of Bacteroidota, Verrucomicrobiota, Myxococcota, and Bdellovibrionota (Table A7). Bacteroidota and Verrucomicrobiota are phyla known for degrading complex organic matter, such as proteins and carbohydrates, which may provide the developing embryo with important nutritional requirements (He et al., 2017; Thomas et al., 2011). Bacteroidota has been found to possibly contribute to the formation of chicken embryonic intestinal microbiota during egg development (Ding et al., 2022), and may be expected to play a similar role in sea turtles. Myxococcota and Bdellovibrionota are phyla known for their predatory lifestyles. Myxococcota can secrete diverse secondary metabolites as antimicrobial proteins and metabolites, which are presumed to aid them in predating a broad range of microbes, including bacteria and fungi (Furness et al., 2020). Bdellovibrionota, on the other hand, is an obligate Gram-negative predator (Li et al., 2021). In both cases, the higher abundance of predatory phyla may protect the turtle embryos from potential pathogens.

At the genus level, hatched eggs were correlated with higher abundances of Sphingopyxis, Pseudonocardia, Devosia, and Cohnella (Table A7). Sphingopyxis and Devosia are genera known for their degradation capabilities. Sphingopyxis have the potential to degrade a number of xenobiotics and other environmental contaminants, which helps them interact and survive in extreme environments (Sharma et al., 2021). Detoxification and degradation of organic pollutants have been identified as dominant functions of the genus Devosia (Talwar et al., 2020). The presence of bacteria capable of degrading environmental contaminants may protect the turtle embryos from being harmed during the incubation process. Pseudonocardia produce diverse secondary metabolites with antimicrobial bioactivities but are mostly known for their symbiotic relationship with fungus-growing ants in which they inhibit entomopathogens that infect the ants (Goldstein & Klassen, 2020). The association with sea turtle eggs may play a role in protecting the eggs from fungal pathogens, such as

Fusarium spp., which have been previously found in unhatched eggs (Brofft Bailey et al., 2018; Sarmiento-Ramírez et al., 2010, 2014). *Cohnella* is a genus known for its cellulolytic or xylanolytic activities (Arneodo et al., 2019), whose relevance to sea turtle egg survival is yet to be determined. We posit that these compounds may be used to reduce root invasion into the eggshell from dune vegetation, which has been found to reduce hatching success in leatherback sea turtles (Conrad et al., 2011).

4.3 | Unhatched egg microbiota—The putative "unhealthy microbiota"

Unhatched eggs were found to have a high species richness (>500 ASVs) within their microbiota, representing more variable bacterial assemblages than hatched eggs. At the phylum level, unhatched (presumed unhealthy) eggs were correlated with higher abundances of Pseudomonadota and Bacillota (Table A7). Pseudomonadota and Bacillota have previously been associated with disease suppression within the rhizosphere and have been previously identified in Fusarium-infected sea turtle eggs (Mendes et al., 2011; Sarmiento-Ramírez et al., 2014). The presence of certain Pseudomonadota (Betaproteobacteria, Gammaproteobacteria) has been suggested as a diagnostic for dysbiosis in humans due to their role in protein degradation and utilization of sugar and oxygen in the gut (Shin et al., 2015). Pseudomonadota has been previously suggested as an indicator of dysbiosis in sea turtles (Campos et al., 2018; Samuelson et al., 2020). Within the present study at the class level for Pseudomonadota, Gammaproteobacteria was found in higher abundance in unhatched eggs, while Alphaproteobacteria was found in higher abundance in hatched eggs (Table A7). Based on our results, higher relative abundances of Gammaproteobacteria could be indicative of dysbiosis within sea turtle egg microbiota. The ratio of Bacillota ("Firmicutes") to Bacteroidota ("Bacteroidetes") (F/B ratio) is

WILEY_MicrobiologyOpen

widely used to track human intestinal homeostasis, with higher ratios being associated with dysbiosis (Abenavoli et al., 2019). The present study also showed a significant difference in the F/B ratio between hatched and unhatched eggs (two-tailed paired samples Wilcoxon test: p = 0.006). Unhatched eggs were determined to have a significantly higher F/B ratio (average F/B ratio = 4.20) than hatched eggs (average F/B ratio = 0.71) (one-tailed paired samples Wilcoxon test: p = 0.003). Further investigation into the F/B ratio in sea turtle eggs should be conducted as it could potentially be used to determine varying levels of dysbiosis within unhatched eggs.

At the genus level, unhatched eggs were correlated with higher abundances of Pseudomonas (Table A7). The genus Pseudomonas was identified as the major driver of the observed differences between hatched and unhatched egg samples, with a higher average relative abundance found in unhatched eggs (19.29%; variance = 7.34%) than in hatched eggs (1.10%; variance = 0.01%) (Figure 4 and Table A7). Pseudomonas spp. have been suggested to play a strong role in sea turtle egg hatching failure (Capri et al., 2023) and have been previously isolated and identified in unhatched sea turtle eggs from other studies (Awong-Taylor et al., 2008; Craven et al., 2007; Keene, 2012; Wyneken et al., 1988). The genus Pseudomonas contains over 200 known taxa (Girard et al., 2021) that have high metabolic diversity, simple nutritional requirements, and wide temperature growth ranges (4°C-42°C) allowing them to inhabit a variety of environments (e.g., soil, water, plants, and association with larger organisms) as commensals and opportunistic pathogens (Chakravarty & Anderson, 2015). The diversity and adaptability of Pseudomonas may help explain why the taxon occurs widely in many environments and was found abundantly in unhatched eggs; however, we cannot be sure if the taxon caused the eggs to not hatch. Pseudomonas spp. have long been associated with proteinaceous food (i.e., eggs, milk, meat) spoilage under aerobic conditions (Raposo et al., 2016), so the presence of Pseudomonas may be due to the decomposition of the unhatched eggs. Future work could be done to try to establish the timeline in which Pseudomonas spp. enter the eggs to determine its functional role in sea turtle eggs.

Alcaligenes and Achromobacter were also found in higher abundance in unhatched eggs than hatched eggs (Figure 4). Both Alcaligenes and Achromobacter are known as opportunistic pathogens and are commonly found in soil, water, and intestinal tracts of vertebrates, such as loggerhead sea turtles (Busse & Stolz, 2006; Trotta et al., 2021). Alcaligenes spp. and Achromobacter spp. have recently been identified in sea turtle eggs, specifically green turtles (Candan & Candan, 2020), our study provides additional support to their presence and provides new insights into their higher abundances in unhatched eggs.

4.4 | Beach environment influence and tracking potential transmission source

The transmission of pathogens into sea turtle eggs is unknown, but we can hypothesize that transmission occurs either through (i)

maternal transmission during the two-week formation period within the uterine tube and oviposition (Funkhouser & Bordenstein, 2013) or (ii) environmental transmission from sand surrounding the nest during the two-month incubation period (Craven et al., 2007). Our study found that 44%-49% of unhatched eggs and 61%-90% of hatched egg microbiota derived from the nest sand microbiota, suggesting that the nest sand environment played a stronger role in shaping both egg microbiota than the cloaca, which was estimated to contribute less than 1% to both hatched and unhatched egg microbiota (Figure 6). Our results support the findings of Capri et al. (2023), who suggest that nest sand microbiota play a crucial role in shaping egg microbiota, with the different bacterial phyla present at each nest potentially being responsible for egg hatching failure. Due to the timing of egg sample collections (after incubation), we cannot dismiss maternal transfer as a critical component of sea turtle egg bacteria acquisition. Maternal transfer of bacteria may play a more important role during the initial stages of embryonic development within the female.

Maternal and environmental transmission, however, are not necessarily mutually exclusive. Sea turtle egg microbiota may potentially stem from a "starter" community from the nesting female, but over time, succession of environmental microbes may occur, depending on the original community composition established maternally, its resilience to change, and random, chance effects. Our study found that sea turtle nests host a different bacterial community than the original beach sand with only 46.55% of the nest microbiota being sourced from the control sand (Figure 2). We hypothesize that the distinction between the two communities may be a result of the mucus excreted by the female during oviposition. The potential antimicrobial properties, warm nest temperature (28.5°C-33.9°C in our study, Table 1), and added nutrients from the female mucus during oviposition may prime the nest sand and eggs to allow for the "starter" microbial community to colonize first during the initial stages of the incubation period. This would allow for a new microbial community to develop within the nest in place of the original beach microbial community. However, the cloaca was found to only account for 0.89% of bacterial introduction in nest sand samples suggesting that the change in nest sand composition from control sand cannot be directly attributed to the female cloaca. The "starter" microbial community within the nest from cloacal mucus may shift from the succession of environmental microbes over time due to the incubation length of sea turtle eggs (40-60 days).

There is also the potential for pathogens to be transmitted by multiple routes (Bright & Bulgheresi, 2010). The exact route of transmission for a pathogen may depend on the direct trade-offs or indirect fitness effects caused by the different routes, the environment the pathogen is in, or based on symbiotic relationships with other microbes (Antonovics et al., 2017; Russell, 2019). Therefore, the likelihood of mixed-mode transmission being used in sea turtle eggs appears high. For example, *Pseudomonas*, which was the most significant driver of differences between hatched and unhatched eggs, was present in all but three samples. Additionally, *Pseudomonas* spp. were found in low abundances in both cloaca (loggerhead =

-WILEY

0.21%, green turtle = 1.02%) and sand samples (loggerhead = 0.48% [control sand] and 1.32% [nest sand], green turtle = 0.81% [control sand] and 0.94% [nest sand]), making the exact route of transmission difficult to isolate. We hypothesize that pathogens may enter the eggs using a mixed-mode route for sea turtle eggs due to the varying conditions of the female reproductive tract and final nesting locations.

We found that 28%–37% of hatched egg and 24%–35% of unhatched egg ASVs did not match the cloaca, control sand, or nest sand (Figure 5) and 1%–4% of hatched egg and 24%–48% of unhatched egg microbiota were determined to be from unknown sources (Figure 6). Therefore, there is still the potential for many other sources of transmission (e.g., water [rain or ocean], vegetation [dune or *Sargassum*], marine debris, and sand macroorganisms) due to the different location of each nest or additional sources before oviposition (e.g., paternal or other maternal sources [i.e., not the cloaca]). Different sea turtle nesting beaches could also be exposed to unique wildlife and vegetation or have different proximities to waste outflows and harbors that can impact nest microbiota.

We found that geographic location alone did not shape sea turtle egg microbiota, however, different combinations of environmental factors between turtle species and beaches explained the bacterial differences in egg microbiota (Tables A9 and A10). The variation in environmental factors between turtle species may be attributed to nest site selection differences between loggerhead and green turtles. Green turtle samples were found closer to dunes (average green turtle nest dune distance = 0.9 m [± 2.2 m], loggerhead = 60.1 m [± 66.0 m]), while loggerhead samples were closer to the high tide line (average loggerhead nest high tide distance = 9.8 m [± 9.1 m], green turtle = 16.9 m [± 10.2 m]; Table 1).

Beach composition, specifically beach width, may also have affected nest site selection and may also explain the variation in environmental factors between beaches. Hillsboro beaches are known for having narrower littoral zones thus restricting the dispersal of microbes and having a greater amount of vegetation and less human activity due to it being a residential beach whereas Fort Lauderdale is more commercialized with wider beaches and less vegetation. Nests were laid further from both the high tide line and dunes in Fort Lauderdale (average distance from high tide line = 13.5 m [±13.0 m], dune = 89.7 m [±64.9 m]) than in Hillsboro (average distance from high tide line = 9.8 m [±4.2 m], dune = 3.1 m [±3.3 m]; Table 1). Dune distance was found to play the most significant role in shaping the bacterial differences in beach sand composition in BIO-ENV analyses. Nest distance from the high tide line was potentially not isolated as the best environmental factor in the BIO-ENV analysis because the tidal distance is constantly changing, while the dune distance is a more permanent variable. However, high tide distance was found to significantly affect the bacterial richness and Shannon diversity in the GLMM analyses (Table A9).

Our hypothesis that the local beach environment has the largest effect in shaping sea turtle egg microbiota rather than the nesting mother's microbiota counters previous research that vertical (parent-

offspring) symbiont transmission increases in terrestrial environments and decreases in aquatic environments (Russell, 2019). Sea turtle nests primarily occur in littoral zones (the average distances of all 2021 sampled nests from the high tide line were 11.6 m [±9.7 m] and 44.8 m [±62.5 m] from the dunes; Table 1); thus, the nests have the opportunity to be exposed to periodic tidal washover and aquatic microbiota. Horizontal (environment-offspring) transmission of aquatic bacteria may increase with nests that are near the high tide line due to increased washover occurrences or from the introduction of pathogens from terrestrial run-off. Nests that are laid near the high tide line have the potential to also be introduced to high levels of fecal indicator bacteria, which have been found to exceed set colonyforming unit levels 2.475 times more during high tide than low tide on South Florida beaches (Aranda et al., 2016). The beach environment, particularly dune distance, was found to play a role in shaping the bacterial community differences within control sand samples. Similarly, nest sand communities can potentially be affected by the beach environment and cause egg bacterial differences, such as species richness and diversity, as we saw in our analyses (Table A9).

Differences in control sand microbiota between beaches were found to be mainly caused by the genera Sphingobacterium, Ochrobactrum, Achromobacter, and Nitratireductor, which all occurred in higher abundances in Fort Lauderdale than Hillsboro. Additionally. Sphingobacterium, Ochrobactrum, and Nitratireductor were all found in higher abundances in hatched eggs than unhatched eggs on both beaches (Figure 4). This may explain why Fort Lauderdale had a significantly higher average hatching success rate (88.78%) when compared to Hillsboro (82.62%) overall in the 2021 nesting season for loggerheads (nonparametric two-tailed t-test: W = 323344. p < 0.001). Our findings suggest that the localized environmental conditions and bacteria at each nest may play an important role in shaping the egg microbiota and potentially affecting their hatching success, as previously suggested (Capri et al., 2023). Further research should be conducted to understand the role these bacteria play in egg-hatching success and their use in beach ecosystem dynamics.

5 | CONCLUSIONS

Studying sea turtle egg microbiomes and their transmission can aid in better conservation methods and management that will then aid in population recovery. Our study produced new insights into the bacterial communities associated with hatched and unhatched eggs in loggerhead and green turtle nests and was the first to find that the environment plays a stronger role in shaping sea turtle egg microbiota than the nesting female. Future research should be conducted to further corroborate these results and determine ways to monitor important sea turtle nesting locations for potentially pathogenic bacteria. Additionally, similar research can be conducted on cloaca, sand, and egg (unhatched and hatched) samples but with the inclusion of mycobiome analyses with the current bacterial analyses. By not including the mycobiome in our current study, a key element VILEY_MicrobiologyOpen

of variation in microbiome-associated hatchability (e.g., *Pseudomonas* associated with fungal growth in unhatched eggs) could be missing and should be further investigated.

Since we found that nest sand is a main source of bacterial transmission for sea turtle eggs, further research should be conducted where the nest's microbiota and environmental parameters (e.g., temperature, gas diffusion and presence, and moisture levels) are analyzed continuously throughout the eggs' incubation period. This would provide better insight into what environmental parameters cause the inundation of pathogens and see if there is a particular stage during the incubation period that it occurs. Additionally, mechanical beach cleaning and beach renourishment projects should be monitored on critical sea turtle nesting beaches to determine if these human activities are altering the sea turtle nesting microbiota and bacterial loads.

Knowing which bacteria are potentially pathogenic to sea turtle eggs could provide additional details to allow for targeted beach monitoring to promote greater hatching success for loggerhead and green turtles in South Florida. This information can also be used to help mitigate human impacts on pathogen transmission into sea turtle eggs. Ideal sea turtle nest relocation zones can be identified on important sea turtle nesting beaches by analyzing the sand microbiota for bacteria correlated with higher hatching success rates and healthy sea turtle eggs. Additionally, sand that is being used for beach renourishment projects can be analyzed for any potentially harmful bacteria before placement in important sea turtle nesting areas to reduce the chance of introducing pathogenic bacteria to sea turtle nests. Sea turtle conservation depends on healthy nesting and hatching, identifying the bacteria influencing the success of sea turtle eggs and understanding their transmission can help reduce threats to their conservation and help protect and preserve these endangered species.

AUTHOR CONTRIBUTIONS

Colleen M. McMaken: data curation (lead), formal analysis (lead), funding acquisition (equal), methodology (lead), visualization (lead), writing—original draft (lead), writing—review & editing (equal); Derek A. Burkholder: conceptualization (equal), investigation (supporting), methodology (supporting), writing—review & editing (equal); Rosanna J. Milligan: data curation (supporting), formal analysis (supporting), visualization (supporting), writing— review & editing (equal); Jose V. Lopez: conceptualization (equal), data curation (supporting), formal analysis (supporting), funding acquisition (equal), investigation (supporting), methodology (supporting), writing—review & editing (equal).

ACKNOWLEDGMENTS

This research was approved under the Marine Turtle Permit #255 (MTP-20-255B & MTP-21-255) by the Florida Fish and Wildlife Conservation Commission (FWC). Funding was provided by the President's Faculty Research and Development Grant (PFRDG; Grant #334882) from Nova Southeastern University. We thank the management and staff of the Broward County Sea Turtle Conservation Program and members of the Conservation, Movement, and

Ecosystem Dynamics (CMED) lab at Nova Southeastern University for assistance in sample collection and nest monitoring. We also

thank members of Dr. Jose V. Lopez's Microbiology and Genetics Laboratory at Nova Southeastern University for their assistance in sample preparation, sequencing, and input on data analysis.

CONFLICT OF INTEREST STATEMENT

The authors declare no conflicts of interest.

DATA AVAILABILITY STATEMENT

Final sequence data were uploaded to NCBI's Short Read Archive (BioProject accession: PRJNA804903: https://www.ncbi.nlm.nih. gov/bioproject/PRJNA804903. Full data files (lists for potential pathogens, shared and unique ASVs, SIMPER analyses, Krona plots) along with the R scripts and files used for this analysis can be found at https://github.com/colmcmaken/Turtle-Egg-Microbiome.

ETHICS STATEMENT

All prevailing local, national, and international regulations and conventions, and normal scientific ethical practices, have been respected. Sampling was authorized by the Florida Fish and Wildlife Conservation Commission (FWC) under Marine Turtle Permit (MTP) #255.

ORCID

Colleen M. McMaken D http://orcid.org/0000-0002-3561-8586

REFERENCES

- Abenavoli, L., Scarpellini, E., Colica, C., Boccuto, L., Salehi, B., Sharifi-Rad, J., Aiello, V., Romano, B., De Lorenzo, A., Izzo, A. A., & Capasso, R. (2019). Gut microbiota and obesity: A role for probiotics. *Nutrients*, 11, 2690.
- Ackerman, R. A. (1996). The nest environment and the embryonic development of sea turtles, *The Biology of Sea Turtles*. CRC Press.
- Al-Bahry, S., Mahmoud, I., Elshafie, A., Al-Harthy, A., Al-Ghafri, S., Al-Amri, I., & Alkindi, A. (2009). Bacterial flora and antibiotic resistance from eggs of green turtles *Chelonia mydas*: An indication of polluted effluents. *Marine Pollution Bulletin*, 58, 720–725.
- Anderson, M. J., Gorley, R. N., & Clarke, K. R. (2008). PERMANOVA+ for PRIMER: Guide to Software and Statistical Methods. PRIMER-E.
- Anderson, M. J., & Willis, T. J. (2003). Canonical analysis of principal coordinates: A useful method of constrained ordination for ecology. *Ecology*, 84, 511–525.
- Antonovics, J., Wilson, A. J., Forbes, M. R., Hauffe, H. C., Kallio, E. R., Leggett, H. C., Longdon, B., Okamura, B., Sait, S. M., & Webster, J. P. (2017). The evolution of transmission mode. *Philosophical Transactions of the Royal Society, B: Biological Sciences*, 372, 20160083.
- Aranda, D., Lopez, J. V., Solo-Gabriele, H. M., & Fleisher, J. M. (2016). Using probabilities of *Enterococci exceedance* and logistic regression to evaluate long term weekly beach monitoring data. *Journal of Water and Health*, 14, 81–89.
- Arneodo, J. D., Etcheverry, C., Thebe, T., Babalola, O. O., Godoy, M. C., & Talia, P. (2019). Molecular evidence that cellulolytic bacterial genus *Cohnella* is widespread among neotropical nasutitermitinae from NE Argentina. *Revista Argentina de microbiología*, 51, 77–80.
- Awong-Taylor, J., Craven, K. S., Griffiths, L., Bass, C., & Muscarella, M. (2008). Comparison of biochemical and molecular methods for the

MicrobiologyOpen

17 of 34

identification of bacterial isolates associated with failed loggerhead sea turtle eggs. *Journal of Applied Microbiology*, 104, 1244–1251.

- Balazs, G. H. (1980). Synopsis of biological data on the green turtle in the Hawaiian Islands. In: U. S. D. O. COMMERCE, N. O. A. A. ADMINISTRATION & N. M. F. SERVICE, (Eds). NOAA Technical Memorandum NMFS.
- Bates, D., Mächler, M., Bolker, B., & Walker, S. (2015). Fitting linear mixedeffects models using Ime4, Journal of Statistical Software 67, 1–48.
- Bézy, V., Hill-Spanik, K., & Plante, C. (2020). Unearthing the sand microbiome of sea turtle nests with disparate survivorship at a massnesting beach in Costa Rica. Aquatic Microbial Ecology, 85, 71–83.
- Bézy, V. S., Valverde, R. A., & Plante, C. J. (2015). Olive ridley sea turtle hatching success as a function of the microbial abundance in nest sand at Ostional, Costa Rica. PLOS ONE, 10, e0118579.
- Boehm, A. B., Yamahara, K. M., & Sassoubre, L. M. (2014). Diversity and transport of microorganisms in intertidal sands of the California coast. Applied and Environmental Microbiology, 80, 3943–3951.
- Bokulich, N. A., Kaehler, B. D., Rideout, J. R., Dillon, M., Bolyen, E., Knight, R., Huttley, G. A., & Gregory Caporaso, J. (2018). Optimizing taxonomic classification of marker-gene amplicon sequences with QIIME 2's q2-feature-classifier plugin. *Microbiome*, *6*, 90.
- Bolyen, E., Rideout, J. R., Dillon, M. R., Bokulich, N. A., Abnet, C. C., Al-Ghalith, G. A., Alexander, H., Alm, E. J., Arumugam, M., Asnicar, F., Bai, Y., Bisanz, J. E., Bittinger, K., Brejnrod, A., Brislawn, C. J., Brown, C. T., Callahan, B. J., Caraballo-Rodríguez, A. M., Chase, J., ... Caporaso, J. G. (2019). Reproducible, interactive, scalable and extensible microbiome data science using QIIME 2. Nature Biotechnology, 37, 852–857.
- Bright, M., & Bulgheresi, S. (2010). A complex journey: Transmission of microbial symbionts. *Nature Reviews Microbiology*, 8, 218–230.
- Brofft Bailey, J., Lamb, M., Walker, M., Weed, C., & Stephenson Craven, K. (2018). Detection of potential fungal pathogens Fusarium falciforme and F. keratoplasticum in unhatched loggerhead turtle eggs using a molecular approach. Endangered Species Research, 36, 111–119.
- Bunte, K., & Abt, S. R. (2001). Particle analysis. Sampling surface and subsurface particle-size distributions in wadable gravel-and cobblebed streams for analyses in sediment transport, hydraulics, and streambed monitoring. US Department of Agriculture, Forest Service, Rocky Mountain Research Station.
- Burkholder, D., Slagle, C., Goodwin, G., & Nease, A. (2020). Broward County Sea Turtle Conservation Program 2020 Technical Report. Nova Southeastern University.
- Busse, H.-J., & Stolz, A. (2006). Achromobacter, Alcaligenes and Related Genera. Springer.
- Butler, K. R. (1997). Coastal protection of sea turtles in Florida. Journal of Land Use & Environmental Law, 13, 399–441.
- Callahan, B. J., McMurdie, P. J., Rosen, M. J., Han, A. W., Johnson, A. J. A., & Holmes, S. P. (2016). DADA2: High-resolution sample inference from Illumina amplicon data. *Nature Methods*, 13, 581–583.
- Campos, P., Guivernau, M., Prenafeta-Boldú, F. X., & Cardona, L. (2018). Fast acquisition of a polysaccharide fermenting gut microbiome by juvenile green turtles *Chelonia mydas* after settlement in coastal habitats. *Microbiome*, 6, 69.
- Candan, E. D. (2018). Molecular identification of fungal isolates and hatching success of green turtle (*Chelonia mydas*) nests. Archives of Microbiology, 200, 911–919.
- Candan, O., & Candan, E. D. (2020). Bacterial diversity of the green turtle (Chelonia mydas) nest environment. Science of the Total Environment, 720, 137717.
- Caporaso, J. G., Lauber, C. L., Walters, W. A., Berg-Lyons, D., Lozupone, C. A., Turnbaugh, P. J., Fierer, N., & Knight, R. (2011). Global patterns of 16S rRNA diversity at a depth of millions of sequences per sample. *Proceedings of the National Academy of Sciences*, 108, 4516–4522.

Capri, F. C., Prazzi, E., Casamento, G., Gambino, D., Cassata, G., & Alduina, R. (2023). Correlation between microbial community and hatching failure in loggerhead sea turtle *Caretta caretta*. *Microbial Ecology*. https://doi.org/10.1007/s00248-023-02197-8

-WILEY

- Carr, A., & Hirth, H. (1961). Social facilitation in green turtle siblings. Animal Behaviour, 9, 68–70.
- Carr, A. F. (1967). So Excellent a Fishe: A Natural History of Sea Turtles. American Museum of Natural History.
- Chakravarty, S., & Anderson, G. 2015. The Genus Pseudomonas. Practical Handbook of Microbiology. Third ed.
- Chan, E.-H., & Solomon, S. E. (1989). The structure and function of the eggshell of the leatherback turtle (*Dermochelys coriacea*) from Malaysia, with notes on attached fungal forms. *Animal Technology*, 40, 91–102.
- Clarke, K. R. (1993). Non-parametric multivariate analyses of changes in community structure. *Austral Ecology*, *18*, 117–143.
- Clarke, K., & Ainsworth, M. (1993). A method of linking multivariate community structure to environmental variables. *Marine Ecology Progress Series*, 92, 205–219. https://doi.org/10.3354/meps092205
- Clarke, K. R., & Gorley, R. N. 2015. PRIMER v7: User Manual/Tutorial Plymouth PRIMER-E Ltd.
- Conrad, J., Wyneken, J., Garner, J., & Garner, S. (2011). Experimental study of dune vegetation impact and control on leatherback sea turtle *Dermochelys coriacea* nests. *Endangered Species Research*, 15, 13–27.
- Craven, K. S., Awong-Taylor, J., Griffiths, L., Bass, C., & Muscarella, M. (2007). Identification of bacterial isolates from unhatched loggerhead (*Caretta caretta*) sea turtle eggs in Georgia, USA. *Marine Turtle Newsletter*, 115, 9–11.
- Davis, N. M., Proctor, D. M., Holmes, S. P., Relman, D. A., & Callahan, B. J. (2018). Simple statistical identification and removal of contaminant sequences in marker-gene and metagenomics data. *Microbiome*, 6, 226.
- De Reu, K., Grijspeerdt, K., Messens, W., Heyndrickx, M., Uyttendaele, M., Debevere, J., & Herman, L. (2006). Eggshell factors influencing eggshell penetration and whole egg contamination by different bacteria, including Salmonella enteritidis. International Journal of Food Microbiology, 112, 253–260.
- Ding, P., Liu, H., Tong, Y., He, X., Yin, X., Yin, Y., Zhang, H., & Song, Z. (2022). Developmental change of yolk microbiota and its role on early colonization of intestinal microbiota in chicken embryo. *Animals: An Open Access Journal from MDPI*, 12, 16.
- Dusa, A. 2021. venn: Draw Venn Diagrams. R package version 1.10.
- Easson, C. G., & Lopez, J. V. (2019). Depth-dependent environmental drivers of microbial plankton community structure in the Northern Gulf of Mexico. Frontiers in Microbiology, 9, 3175.
- Fournier, J., Gallon, R. K., & Paris, R. (2014). G2Sd: A new R package for the statistical analysis of unconsolidated sediments, Géomorphologie: relief, processus, environnement 20, 73–78.
- Funkhouser, L. J., & Bordenstein, S. R. (2013). Mom knows best: The universality of maternal microbial transmission. PLOS Biology, 11, e1001631.
- Furness, E., Whitworth, D. E., & Zwarycz, A. (2020). Predatory Interactions Between Myxobacteria and Their Prey. In: E. Jurkevitch & R. J. Mitchell (Eds.), *The Ecology of Predation at the Microscale*. Springer International Publishing.
- Gifari, T., Elfidasari, D., & Sugoro, I. (2018). The effects of contaminant microorganism towards *Chelonia mydas* eggs hatchery results in Pangumbahan green sea turtles conservation, Sukabumi, Indonesia. *Biodiversitas Journal of Biological Diversity*, 19, 1207–1212.
- Girard, L., Lood, C., Höfte, M., Vandamme, P., Rokni-Zadeh, H., van Noort, V., Lavigne, R., & De Mot, R. (2021). The everexpanding *Pseudomonas* genus: Description of 43 new species and partition of the *Pseudomonas putida* group. *Microorganisms*, 9, 1766.

WILEY_MicrobiologyOpen

18 of 34

- Gleason, F. H., Allerstorfer, M., & Lilje, O. (2020). Newly emerging diseases of marine turtles, especially sea turtle egg fusariosis (SEFT), caused by species in the *Fusarium solani* complex (FSSC). *Mycology*, 11, 184–194.
- Goldstein, S. L., & Klassen, J. L. (2020). Pseudonocardia Symbionts of fungus-growing ants and the evolution of defensive secondary metabolism. Frontiers in Microbiology, 11, 621041.
- Greenwald, W. W., Li, H., Benaglio, P., Jakubosky, D., Matsui, H., Schmitt, A., Selvaraj, S., D'Antonio, M., D'antonio-Chronowska, A., Smith, E. N., & Frazer, K. A. (2019). Subtle changes in chromatin loop contact propensity are associated with differential gene regulation and expression. *Nature Communications*, 10, 1054.
- He, S., Stevens, S. L. R., Chan, L.-K., Bertilsson, S., Glavina del Rio, T., Tringe, S. G., Malmstrom, R. R., & McMahon, K. D. (2017). Ecophysiology of freshwater verrucomicrobia inferred from Metagenome-assembled genomes. *mSphere*, 2, e00277–17.
- Hoh, D. Z., Lin, Y.-F., Liu, W.-A., Sidique, S. N. M., & Tsai, I. J. (2020). Nest microbiota and pathogen abundance in sea turtle hatcheries. *Fungal Ecology*, 47, 100964.
- Honarvar, S., Spotila, J. R., & O'Connor, M. P. (2011). Microbial community structure in sand on two olive ridley arribada nesting beaches, Playa La Flor, Nicaragua and Playa Nancite, Costa Rica. *Journal of Experimental Marine Biology and Ecology*, 409, 339–344.
- Hsieh, T. C., Ma, K. H., & Chao, A. (2016). iNEXT: An R package for rarefaction and extrapolation of species diversity (Hill numbers). *Methods in Ecology and Evolution*, 7, 1451–1456.
- Hu, A., Wang, H., Cao, M., Rashid, A., Li, M., & Yu, C.-P. (2019). Environmental filtering drives the assembly of habitat generalists and specialists in the coastal sand microbial communities of Southern China. *Microorganisms*, 7, 598.
- Jackson, J. B. C. (1997). Reefs since Columbus. Coral Reefs, 16, S23-S32.
- Katoh, K., Misawa, K., Kuma, K., & Miyata, T. (2002). MAFFT: A novel method for rapid multiple sequence alignment based on fast Fourier transform. Nucleic Acids Research, 30, 3059–3066.
- Keene, E. L. (2012). Microorganisms from sand, cloacal fluid, and eggs of Lepidochelys olivacea and standard testing of cloacal fluid antimicrobial properties. Master of Science, Purdue University.
- Knight, R., Jansson, J., Field, D., Fierer, N., Desai, N., Fuhrman, J. A., Hugenholtz, P., van der Lelie, D., Meyer, F., Stevens, R., Bailey, M. J., Gordon, J. I., Kowalchuk, G. A., & Gilbert, J. A. (2012). Unlocking the potential of metagenomics through replicated experimental design. *Nature Biotechnology*, 30, 513–520.
- Knights, D., Kuczynski, J., Charlson, E. S., Zaneveld, J., Mozer, M. C., Collman, R. G., Bushman, F. D., Knight, R., & Kelley, S. T. (2011). Bayesian community-wide culture-independent microbial source tracking. *Nature Methods*, *8*, 761–763.
- Lahti, L., & Shetty, S. 2012-2019. microbiome R package.
- Lenth, R. V. 2023. emmeans: Estimated Marginal Means, aka Least-Squares Means. 1.8.5 Ed.
- Li, Q.-M., Zhou, Y.-L., Wei, Z.-F., & Wang, Y. (2021). Phylogenomic insights into distribution and adaptation of bdellovibrionota in marine waters. *Microorganisms*, *9*, 757.
- Mantel, N., & Valand, R. S. (1970). A technique of nonparametric multivariate analysis. *Biometrics*, 26, 547–558.
- Margalef, R. (1958). Temporal succession and spatial heterogeneity in phytoplankton. In: A. A. Buzzati-traverso, (Ed.), *Perspectives in Marine Biology*. University of California Press.
- McFall-Ngai, M., Hadfield, M. G., Bosch, T. C. G., Carey, H. V., Domazet-Lošo, T., Douglas, A. E., Dubilier, N., Eberl, G., Fukami, T., Gilbert, S. F., Hentschel, U., King, N., Kjelleberg, S., Knoll, A. H., Kremer, N., Mazmanian, S. K., Metcalf, J. L., Nealson, K., Pierce, N. E., ... Wernegreen, J. J. (2013). Animals in a bacterial world, a new imperative for the life sciences. *Proceedings of the National Academy* of Sciences, 110, 3229–3236.

- McMurdie, P. J., & Holmes, S. (2013). phyloseq: An R package for reproducible interactive analysis and graphics of microbiome census data. PLOS ONE, 8, e61217.
- Mendes, R., Kruijt, M., de Bruijn, I., Dekkers, E., van der Voort, M., Schneider, J. H. M., Piceno, Y. M., DeSantis, T. Z., Andersen, G. L., Bakker, P. A. H. M., & Raaijmakers, J. M. (2011). Deciphering the rhizosphere microbiome for disease-suppressive bacteria. *Science*, 332, 1097–1100.
- Miller, J. D. (1985). Embryology of Marine Turtles. In: C. Gans, F. Billett & P. F. A. Maderson, (Eds.), *Biology of the Reptilia*. John Wiley & Sons.
- Mortimer, J. A. (1990). The influence of beach sand characteristics on the nesting behavior and clutch survival of green turtles (*Chelonia mydas*). Copeia, 1990, 802–817.
- Oksanen, J., Blanchet, F. G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., Minchin, P. R., O'Hara, R. B., Simpson, G. L., Solymos, P., Stevens, M. H. H., Szoecs, E., & Wagner, H. 2020. vegan: Community Ecology Package.
- Oren, A., & Garrity, G. M. (2021). Valid publication of the names of fortytwo phyla of prokaryotes. *International Journal of Systematic and Evolutionary Microbiology*, 71, 005056. https://doi.org/10.1099/ ijsem.0.005056
- Owens, D. W. (1980). The comparative reproductive physiology of sea turtles. *American Zoologist*, 20, 549-563.
- Piggot, A. M., Klaus, J. S., Johnson, S., Phillips, M. C., & Solo-Gabriele, H. M. (2012). Relationship between enterococcal levels and sediment biofilms at recreational beaches in South Florida. *Applied and Environmental Microbiology*, *78*, 5973–5982.
- Poiani, A. (2010). Do cloacal pathogenic microbes behave as sexually transmitted parasites in birds? *The Open Ornithology Journal*, *3*, 72–85.
- Praja, R. N., Yudhana, A., Haditanojo, W., & Oktaviana, V. (2021). Short communication: Antimicrobial properties in cloacal fluid of olive ridley sea turtle (*Lepidochelys olivacea*). Biodiversitas Journal of Biological Diversity, 22, 3671–3676.
- Prange, H. D., & Ackerman, R. A. (1974). Oxygen consumption and mechanisms of gas exchange of green turtle (*Chelonia mydas*) eggs and hatchlings. *Copeia*, 1974, 758–763.
- Price, M. N., Dehal, P. S., & Arkin, A. P. (2010). FastTree 2–approximately maximum-likelihood trees for large alignments. PLOS ONE, 5, e9490.
- Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., Peplies, J., & Glöckner, F. O. (2012). The SILVA ribosomal RNA gene database project: Improved data processing and web-based tools. *Nucleic Acids Research*, 41, D590–D596.
- R Core Team. (2020). R: A language and environment for statistical computing. R Foundation for Statistical Computing.
- Raposo, A., Pérez, E., de Faria, C. T., Ferrús, M. A., & Carrascosa, C. (2016). Food Spoilage by *Pseudomonas* spp.—An Overview. In: O. V. Singh, (Ed.), *Foodborne Pathogens and Antibiotic Resistance*.
- Rees, G. N., Baldwin, D. S., Watson, G. O., Perryman, S., & Nielsen, D. L. (2004). Ordination and significance testing of microbial community composition derived from terminal restriction fragment length polymorphisms: Application of multivariate statistics. *Antonie Van Leeuwenhoek*, 86, 339–347.
- Rosado-Rodríguez, G., & Maldonado-Ramírez, S. L. (2016). Mycelial fungal diversity associated with the leatherback sea turtle (*Dermochelys coriacea*) nests from Western Puerto Rico. *Chelonian Conservation and Biology*, 15, 265–272.
- Russell, S. L. (2019). Transmission mode is associated with environment type and taxa across bacteria-eukaryote symbioses: A systematic review and meta-analysis. *FEMS Microbiology Letters*, *366*, fnz013.
- Samuelson, M., Pulis, E., Ray, C., Arias, C., Samuelson, D., Mattson, E., & Solangi, M. (2020). Analysis of the fecal microbiome in Kemp's ridley sea turtles (*Lepidochelys kempii*) undergoing rehabilitation. *Endangered Species Research*, 43, 121–131.

- Sarmiento-Ramírez, J. M., Abella, E., Martín, M. P., Tellería, M. T., López-Jurado, L. F., Marco, A., & Diéguez-Uribeondo, J. (2010). Fusarium solani is responsible for mass mortalities in nests of loggerhead sea turtle, Caretta caretta, in Boavista, Cape Verde. FEMS Microbiology Letters, 312, 192–200.
- Sarmiento-Ramírez, J. M., Abella-Pérez, E., Phillott, A. D., Sim, J., Van West, P., Martín, M. P., Marco, A., & Diéguez-Uribeondo, J. (2014). Global distribution of two fungal pathogens threatening endangered sea turtles. *PLOS ONE*, *9*, e85853.
- Sarmiento-Ramirez, J. M., Sim, J., Van West, P., & Dieguez-Uribeondo, J. (2017). Isolation of fungal pathogens from eggs of the endangered sea turtle species Chelonia mydas in Ascension Island. Journal of the Marine Biological Association of the United Kingdom, 97, 661–667.
- Sarmiento-Ramírez, J. M., van der Voort, M., Raaijmakers, J. M., & Diéguez-Uribeondo, J. (2014). Unravelling the microbiome of eggs of the endangered sea turtle *Eretmochelys imbricata* identifies bacteria with activity against the emerging pathogen *Fusarium falciforme*. *PLoS ONE*, 9, e95206.
- Scheelings, T. F. (2019). The microbiota of sea turtles. Doctor of Philosophy, Monash University.
- Shannon, C. E., & Weaver, W. (1949). The mathematical theory of communication. The University of Illinois Press.
- Sharma, M., Khurana, H., Singh, D. N., & Negi, R. K. (2021). The genus Sphingopyxis: Systematics, ecology, and bioremediation potential - A review. *Journal of Environmental Management*, 280, 111744.
- Shin, N.-R., Whon, T. W., & Bae, J.-W. (2015). Proteobacteria: Microbial signature of dysbiosis in gut microbiota. *Trends in Biotechnology*, 33, 496–503.
- Sidique, S. N. M., Azuddin, N. F., & Joseph, J. (2017). First report of *Fusarium* species at nesting sites of endangered sea turtles in Terengganu and Melaka, Malaysia. *Malaysian Applied Biology*, 46, 195–205.

Simpson, E. H. (1949). Measurement of diversity. Nature, 163, 688.

- Smyth, C. W., Sarmiento-Ramírez, J. M., Short, D. P. G., Diéguez-Uribeondo, J., O'Donnell, K., & Geiser, D. M. (2019). Unraveling the ecology and epidemiology of an emerging fungal disease, sea turtle egg fusariosis (STEF). *PLOS Pathogens*, 15, e1007682.
- Snell, H. L., & Fritts, T. H. (1983). The significance of diurnal terrestrial emergence of green turtles (*Chelonia mydas*) in the Galapagos Archipelago. *Biotropica*, 15, 285–291.
- Soslau, G., Spotila, J. R., Chun, A., Yi, S., & Weber, K. T. (2011). Potentially lethal bacteria in leatherback turtle eggs in the wild threaten both turtles and conservationists. *Journal of Experimental Marine Biology* and Ecology, 410, 101–106.
- Spotila, J. R. (2004). Sea turtles: A complete guide to their biology, behavior, and conservation. Johns Hopkins University Press.
- Staley, C., & Sadowsky, M. J. (2016). Regional similarities and consistent patterns of local variation in beach sand bacterial communities

throughout the northern hemisphere. Applied and Environmental Microbiology, 82, 2751-2762.

WILEY

- Talwar, C., Nagar, S., Kumar, R., Scaria, J., Lal, R., & Negi, R. K. (2020). Defining the environmental adaptations of genus *Devosia*: Insights into its expansive short peptide transport system and positively selected genes. *Scientific Reports*, 10, 1151.
- Thomas, F., Hehemann, J.-H., Rebuffet, E., Czjzek, M., & Michel, G. (2011). Environmental and gut Bacteroidetes: The food connection. *Frontiers* in *Microbiology*, 2, 93.
- Thompson, L. R., Sanders, J. G., McDonald, D., Amir, A., Ladau, J., Locey, K. J., Prill, R. J., Tripathi, A., Gibbons, S. M., Ackermann, G., Navas-Molina, J. A., Janssen, S., Kopylova, E., Vázquez-Baeza, Y., González, A., Morton, J. T., Mirarab, S., Zech Xu, Z., Jiang, L., ... Kerkhof, L. J. (2017). A communal catalogue reveals earth's multiscale microbial diversity. *Nature*, *551*, 457–463.
- Trask, P. D. (1932). Origin and environment of source sediments of petroleum. Gulf Publishing Co.
- Trotta, A., Marinaro, M., Sposato, A., Galgano, M., Ciccarelli, S., Paci, S., & Corrente, M. (2021). Antimicrobial resistance in Loggerhead Sea Turtles (*Caretta caretta*): A comparison between clinical and commensal bacterial isolates. *Animals: An Open Access Journal from MDPI*, 11, 2435.
- Vecchioni, L., Pace, A., Sucato, A., Berlinghieri, F., Cambera, I., Visconti, G., Hochscheid, S., Arculeo, M., & Alduina, R. (2022). Unveiling the egg microbiota of the loggerhead sea turtle *Caretta caretta* in nesting beaches of the Mediterranean Sea. *PLOS ONE*, 17, e0268345.
- Wyneken, J., Burke, T. J., Salmon, M., & Pedersen, D. K. (1988). Egg failure in natural and relocated sea turtle nests. *Journal of Herpetology*, 22, 88–96.
- Yilmaz, P., Parfrey, L. W., Yarza, P., Gerken, J., Pruesse, E., Quast, C., Schweer, T., Peplies, J., Ludwig, W., & Glöckner, F. O. (2014). The SILVA and "all-species living tree project (LTP)" taxonomic frameworks. *Nucleic Acids Research*, 42, D643–D648.
- Zaneveld, J. R., McMinds, R., & Vega Thurber, R. (2017). Stress and stability: Applying the Anna Karenina principle to animal microbiomes. *Nature Microbiology*, *2*, 17121.

How to cite this article: McMaken, C. M., Burkholder, D. A., Milligan, R. J., & Lopez, J. V. (2023). Potential impacts of environmental bacteria on the microbiota of loggerhead (*Caretta caretta*) and green (*Chelonia mydas*) sea turtle eggs and their hatching success. *MicrobiologyOpen*, *12*, e1363. https://doi.org/10.1002/mbo3.1363

APPENDIX

 TABLE A1
 Number of sequence reads and bacterial taxonomy count per sample.

| Sample ID | Nest number | Sample type | Turtle species | Beach | QIIME sequence reads | ASV count |
|--------------|----------------|-------------|----------------|-----------------|----------------------------|--------------|
| C.1166.CM.H | 1166 | Cloaca | Green Turtle | Hillsboro | 100279 | 121 |
| C.1167.CM.H | 1167 | Cloaca | Green Turtle | Hillsboro | 127488 | 113 |
| C.1168.CM.H | 1168 | Cloaca | Green Turtle | Hillsboro | 170528 | 153 |
| C.127.CC.H | 127 | Cloaca | Loggerhead | Hillsboro | 46217 | 268 |
| C.236.CC.H | 236 | Cloaca | Loggerhead | Hillsboro | 123741 | 797 |
| C.242.CC.H | 242 | Cloaca | Loggerhead | Hillsboro | 59700 | 98 |
| C.282.CC.H | 282 | Cloaca | Loggerhead | Hillsboro | 61555 | 506 |
| C.322.CC.H | 322 | Cloaca | Loggerhead | Hillsboro | 103713 | 631 |
| C.323.CC.H | 323 | Cloaca | Loggerhead | Hillsboro | 125246 | 457 |
| C.392.CC.H | 392 | Cloaca | Loggerhead | Hillsboro | 137084 | 424 |
| C.395.CC.H | 395 | Cloaca | Loggerhead | Hillsboro | 90439 | 204 |
| C.456.CC.F | 456 | Cloaca | Loggerhead | Fort Lauderdale | 143942 | 261 |
| C.457.CC.F | 457 | Cloaca | Loggerhead | Fort Lauderdale | 84848 | 451 |
| C.463.CC.F | 463 | Cloaca | Loggerhead | Fort Lauderdale | 114538 | 317 |
| C.463.CC.H | 463 | Cloaca | Loggerhead | Hillsboro | 140709 | 270 |
| C.519.CC.F | 519 | Cloaca | Loggerhead | Fort Lauderdale | 135444 | 243 |
| C.520.CC.F | 520 | Cloaca | Loggerhead | Fort Lauderdale | 51276 | 101 |
| C.521.CC.F | 521 | Cloaca | Loggerhead | Fort Lauderdale | 146274 | 139 |
| C.522.CC.F | 522 | Cloaca | Loggerhead | Fort Lauderdale | 135869 | 164 |
| C.695.CM.H | 695 | Cloaca | Green Turtle | Hillsboro | 113633 | 130 |
| C.696.CM.H | 696 | Cloaca | Green Turtle | Hillsboro | 83683 | 170 |
| C.697.CC.F | 697 | Cloaca | Loggerhead | Fort Lauderdale | 113112 | 242 |
| C.698.CC.F | 698 | Cloaca | Loggerhead | Fort Lauderdale | 132994 | 259 |
| C.699.CC.F | 699 | Cloaca | Loggerhead | Fort Lauderdale | 136480 | 185 |
| C.823.CM.F | 823 | Cloaca | Green Turtle | Fort Lauderdale | 115898 | 167 |
| C.867.CM.F | 867 | Cloaca | Green Turtle | Fort Lauderdale | 130248 | 155 |
| C.892.CC.F | 892 | Cloaca | Loggerhead | Fort Lauderdale | 165918 | 227 |
| H1.1166.CM.H | 1166 | Hatched Egg | Green Turtle | Hillsboro | 126697 | 135 |
| H1.1167.CM.H | 1167 | Hatched Egg | Green Turtle | Hillsboro | 148181 | 192 |
| H1.1168.CM.H | 1168 | Hatched Egg | Green Turtle | Hillsboro | 170681 | 152 |
| H1.127.CC.H | 127 | Hatched Egg | Loggerhead | Hillsboro | 65207 | 157 |
| H1.236.CC.H | 236 | Hatched Egg | Loggerhead | Hillsboro | 125762 | 137 |
| H1.242.CC.H | 242 | Hatched Egg | Loggerhead | Hillsboro | 115605 | 188 |
| H1.282.CC.H | 282 | Hatched Egg | Loggerhead | Hillsboro | 107474 | 244 |

-WILEY

| Sample ID | Nest number | Sample type | Turtle species | Beach | QIIME sequence reads | ASV count |
|--------------|----------------|-------------|----------------|-----------------|----------------------------|--------------|
| H1.322.CC.H | 322 | Hatched Egg | Loggerhead | Hillsboro | 115932 | 151 |
| H1.323.CC.H | 323 | Hatched Egg | Loggerhead | Hillsboro | 71369 | 93 |
| H1.392.CC.H | 392 | Hatched Egg | Loggerhead | Hillsboro | 114411 | 113 |
| H1.395.CC.H | 395 | Hatched Egg | Loggerhead | Hillsboro | 102350 | 104 |
| H1.456.CC.F | 456 | Hatched Egg | Loggerhead | Fort Lauderdale | 127637 | 129 |
| H1.457.CC.F | 457 | Hatched Egg | Loggerhead | Fort Lauderdale | 126113 | 297 |
| H1.463.CC.F | 463 | Hatched Egg | Loggerhead | Fort Lauderdale | 70929 | 92 |
| H1.463.CC.H | 463 | Hatched Egg | Loggerhead | Hillsboro | 149186 | 265 |
| H1.519.CC.F | 519 | Hatched Egg | Loggerhead | Fort Lauderdale | 137968 | 122 |
| H1.520.CC.F | 520 | Hatched Egg | Loggerhead | Fort Lauderdale | 85425 | 87 |
| H1.521.CC.F | 521 | Hatched Egg | Loggerhead | Fort Lauderdale | 77047 | 88 |
| H1.522.CC.F | 522 | Hatched Egg | Loggerhead | Fort Lauderdale | 121755 | 136 |
| H1.695.CM.H | 695 | Hatched Egg | Green Turtle | Hillsboro | 113254 | 140 |
| H1.696.CM.H | 696 | Hatched Egg | Green Turtle | Hillsboro | 87856 | 150 |
| H1.697.CC.F | 697 | Hatched Egg | Loggerhead | Fort Lauderdale | 106900 | 109 |
| H1.698.CC.F | 698 | Hatched Egg | Loggerhead | Fort Lauderdale | 111885 | 143 |
| H1.699.CC.F | 699 | Hatched Egg | Loggerhead | Fort Lauderdale | 88232 | 109 |
| H1.823.CM.F | 823 | Hatched Egg | Green Turtle | Fort Lauderdale | 146423 | 118 |
| H1.867.CM.F | 867 | Hatched Egg | Green Turtle | Fort Lauderdale | 139709 | 132 |
| H1.892.CC.F | 892 | Hatched Egg | Loggerhead | Fort Lauderdale | 109597 | 137 |
| H2.1166.CM.H | 1166 | Hatched Egg | Green Turtle | Hillsboro | 119057 | 131 |
| H2.1167.CM.H | 1167 | Hatched Egg | Green Turtle | Hillsboro | 122004 | 207 |
| H2.1168.CM.H | 1168 | Hatched Egg | Green Turtle | Hillsboro | 160511 | 156 |
| H2.127.CC.H | 127 | Hatched Egg | Loggerhead | Hillsboro | 81923 | 190 |
| H2.236.CC.H | 236 | Hatched Egg | Loggerhead | Hillsboro | 99275 | 193 |
| H2.242.CC.H | 242 | Hatched Egg | Loggerhead | Hillsboro | 83242 | 140 |
| H2.282.CC.H | 282 | Hatched Egg | Loggerhead | Hillsboro | 125318 | 341 |
| H2.322.CC.H | 322 | Hatched Egg | Loggerhead | Hillsboro | 139238 | 135 |
| H2.323.CC.H | 323 | Hatched Egg | Loggerhead | Hillsboro | 93569 | 133 |
| H2.392.CC.H | 392 | Hatched Egg | Loggerhead | Hillsboro | 98448 | 145 |
| H2.395.CC.H | 395 | Hatched Egg | Loggerhead | Hillsboro | 103014 | 94 |
| H2.456.CC.F | 456 | Hatched Egg | Loggerhead | Fort Lauderdale | 140092 | 231 |
| H2.457.CC.F | 457 | Hatched Egg | Loggerhead | Fort Lauderdale | 102299 | 263 |
| H2.463.CC.F | 463 | Hatched Egg | Loggerhead | Fort Lauderdale | 105816 | 154 |
| H2.463.CC.H | 463 | Hatched Egg | Loggerhead | Hillsboro | 190432 | 200 |
| H2.519.CC.F | 519 | Hatched Egg | Loggerhead | Fort Lauderdale | 100963 | 179 |
| H2.520.CC.F | 520 | Hatched Egg | Loggerhead | Fort Lauderdale | 73146 | 85 |

(Continues)

TABLE A1 (Continued)

| Sample ID | Nest number | Sample type | Turtle species | Beach | QIIME sequence reads | ASV count |
|--------------|----------------|-------------|----------------|-----------------|----------------------------|--------------|
| H2.521.CC.F | 521 | Hatched Egg | Loggerhead | Fort Lauderdale | 91415 | 137 |
| H2.522.CC.F | 522 | Hatched Egg | Loggerhead | Fort Lauderdale | 130659 | 144 |
| H2.695.CM.H | 695 | Hatched Egg | Green Turtle | Hillsboro | 96809 | 130 |
| H2.696.CM.H | 696 | Hatched Egg | Green Turtle | Hillsboro | 105326 | 168 |
| H2.697.CC.F | 697 | Hatched Egg | Loggerhead | Fort Lauderdale | 123403 | 144 |
| H2.698.CC.F | 698 | Hatched Egg | Loggerhead | Fort Lauderdale | 137664 | 146 |
| H2.699.CC.F | 699 | Hatched Egg | Loggerhead | Fort Lauderdale | 109190 | 115 |
| H2.823.CM.F | 823 | Hatched Egg | Green Turtle | Fort Lauderdale | 162447 | 116 |
| H2.867.CM.F | 867 | Hatched Egg | Green Turtle | Fort Lauderdale | 140323 | 131 |
| H2.892.CC.F | 892 | Hatched Egg | Loggerhead | Fort Lauderdale | 126565 | 153 |
| H3.1166.CM.H | 1166 | Hatched Egg | Green Turtle | Hillsboro | 113366 | 131 |
| H3.1167.CM.H | 1167 | Hatched Egg | Green Turtle | Hillsboro | 129611 | 180 |
| H3.1168.CM.H | 1168 | Hatched Egg | Green Turtle | Hillsboro | 119282 | 133 |
| H3.127.CC.H | 127 | Hatched Egg | Loggerhead | Hillsboro | 105554 | 178 |
| H3.236.CC.H | 236 | Hatched Egg | Loggerhead | Hillsboro | 96562 | 267 |
| H3.242.CC.H | 242 | Hatched Egg | Loggerhead | Hillsboro | 69680 | 110 |
| H3.282.CC.H | 282 | Hatched Egg | Loggerhead | Hillsboro | 163402 | 409 |
| H3.322.CC.H | 322 | Hatched Egg | Loggerhead | Hillsboro | 113001 | 145 |
| H3.323.CC.H | 323 | Hatched Egg | Loggerhead | Hillsboro | 92665 | 190 |
| H3.392.CC.H | 392 | Hatched Egg | Loggerhead | Hillsboro | 151009 | 200 |
| H3.395.CC.H | 395 | Hatched Egg | Loggerhead | Hillsboro | 103384 | 95 |
| H3.456.CC.F | 456 | Hatched Egg | Loggerhead | Fort Lauderdale | 120150 | 174 |
| H3.457.CC.F | 457 | Hatched Egg | Loggerhead | Fort Lauderdale | 111353 | 212 |
| H3.463.CC.F | 463 | Hatched Egg | Loggerhead | Fort Lauderdale | 110824 | 104 |
| H3.463.CC.H | 463 | Hatched Egg | Loggerhead | Hillsboro | 220501 | 141 |
| H3.519.CC.F | 519 | Hatched Egg | Loggerhead | Fort Lauderdale | 101976 | 112 |
| H3.520.CC.F | 520 | Hatched Egg | Loggerhead | Fort Lauderdale | 83727 | 98 |
| H3.521.CC.F | 521 | Hatched Egg | Loggerhead | Fort Lauderdale | 122011 | 130 |
| H3.522.CC.F | 522 | Hatched Egg | Loggerhead | Fort Lauderdale | 98815 | 120 |
| H3.695.CM.H | 695 | Hatched Egg | Green Turtle | Hillsboro | 96778 | 148 |
| H3.696.CM.H | 696 | Hatched Egg | Green Turtle | Hillsboro | 102833 | 151 |
| H3.697.CC.F | 697 | Hatched Egg | Loggerhead | Fort Lauderdale | 95256 | 118 |
| H3.698.CC.F | 698 | Hatched Egg | Loggerhead | Fort Lauderdale | 150883 | 123 |
| H3.699.CC.F | 699 | Hatched Egg | Loggerhead | Fort Lauderdale | 134604 | 125 |
| H3.823.CM.F | 823 | Hatched Egg | Green Turtle | Fort Lauderdale | 139100 | 117 |
| H3.867.CM.F | 867 | Hatched Egg | Green Turtle | Fort Lauderdale | 111563 | 114 |
| H3.892.CC.F | 892 | Hatched Egg | Loggerhead | Fort Lauderdale | 163214 | 159 |

| Sample ID | Nest number | Sample type | Turtle species | Beach | QIIME sequence reads | ASV count |
|-------------|----------------|--------------|----------------|-----------------|----------------------------|--------------|
| H4.395.CC.H | 395 | Hatched Egg | Loggerhead | Hillsboro | 84968 | 86 |
| H5.395.CC.H | 395 | Hatched Egg | Loggerhead | Hillsboro | 133163 | 145 |
| H6.395.CC.H | 395 | Hatched Egg | Loggerhead | Hillsboro | 135936 | 95 |
| N.1166.CM.H | 1166 | Nest Sand | Green Turtle | Hillsboro | 106640 | 427 |
| N.1167.CM.H | 1167 | Nest Sand | Green Turtle | Hillsboro | 149852 | 520 |
| N.1168.CM.H | 1168 | Nest Sand | Green Turtle | Hillsboro | 145635 | 261 |
| N.127.CC.H | 127 | Nest Sand | Loggerhead | Hillsboro | 116838 | 496 |
| N.236.CC.H | 236 | Nest Sand | Loggerhead | Hillsboro | 103352 | 814 |
| N.242.CC.H | 242 | Nest Sand | Loggerhead | Hillsboro | 128324 | 773 |
| N.282.CC.H | 282 | Nest Sand | Loggerhead | Hillsboro | 114036 | 751 |
| N.322.CC.H | 322 | Nest Sand | Loggerhead | Hillsboro | 127019 | 226 |
| N.323.CC.H | 323 | Nest Sand | Loggerhead | Hillsboro | 94186 | 434 |
| N.392.CC.H | 392 | Nest Sand | Loggerhead | Hillsboro | 174355 | 437 |
| N.395.CC.H | 395 | Nest Sand | Loggerhead | Hillsboro | 185642 | 462 |
| N.456.CC.F | 456 | Nest Sand | Loggerhead | Fort Lauderdale | 127266 | 422 |
| N.457.CC.F | 457 | Nest Sand | Loggerhead | Fort Lauderdale | 146810 | 341 |
| N.463.CC.F | 463 | Nest Sand | Loggerhead | Fort Lauderdale | 143510 | 435 |
| N.463.CC.H | 463 | Nest Sand | Loggerhead | Hillsboro | 123118 | 550 |
| N.519.CC.F | 519 | Nest Sand | Loggerhead | Fort Lauderdale | 134679 | 547 |
| N.520.CC.F | 520 | Nest Sand | Loggerhead | Fort Lauderdale | 71281 | 180 |
| N.521.CC.F | 521 | Nest Sand | Loggerhead | Fort Lauderdale | 170673 | 726 |
| N.522.CC.F | 522 | Nest Sand | Loggerhead | Fort Lauderdale | 73435 | 195 |
| N.695.CM.H | 695 | Nest Sand | Green Turtle | Hillsboro | 106954 | 466 |
| N.696.CM.H | 696 | Nest Sand | Green Turtle | Hillsboro | 130122 | 341 |
| N.697.CC.F | 697 | Nest Sand | Loggerhead | Fort Lauderdale | 107111 | 195 |
| N.698.CC.F | 698 | Nest Sand | Loggerhead | Fort Lauderdale | 159366 | 373 |
| N.699.CC.F | 699 | Nest Sand | Loggerhead | Fort Lauderdale | 108708 | 274 |
| N.823.CM.F | 823 | Nest Sand | Green Turtle | Fort Lauderdale | 169853 | 197 |
| N.867.CM.F | 867 | Nest Sand | Green Turtle | Fort Lauderdale | 152182 | 338 |
| N.892.CC.F | 892 | Nest Sand | Loggerhead | Fort Lauderdale | 139615 | 216 |
| S.1166.CM.H | 1166 | Control Sand | Green Turtle | Hillsboro | 103043 | 212 |
| S.1167.CM.H | 1167 | Control Sand | Green Turtle | Hillsboro | 153048 | 839 |
| S.1168.CM.H | 1168 | Control Sand | Green Turtle | Hillsboro | 115917 | 292 |
| S.127.CC.H | 127 | Control Sand | Loggerhead | Hillsboro | 75476 | 1158 |
| S.236.CC.H | 236 | Control Sand | Loggerhead | Hillsboro | 77044 | 1159 |
| S.242.CC.H | 242 | Control Sand | Loggerhead | Hillsboro | 84761 | 1322 |
| S.282.CC.H | 282 | Control Sand | Loggerhead | Hillsboro | 77507 | 1322 |

(Continues)

TABLE A1 (Continued)

| Sample ID | Nest number | Sample type | Turtle species | Beach | QIIME sequence reads | ASV count |
|--------------|----------------|---------------|----------------|-----------------|----------------------------|--------------|
| S.322.CC.H | 322 | Control Sand | Loggerhead | Hillsboro | 79843 | 1122 |
| S.323.CC.H | 323 | Control Sand | Loggerhead | Hillsboro | 72290 | 1104 |
| S.392.CC.H | 392 | Control Sand | Loggerhead | Hillsboro | 77743 | 1253 |
| S.395.CC.H | 395 | Control Sand | Loggerhead | Hillsboro | 103204 | 1209 |
| S.456.CC.F | 456 | Control Sand | Loggerhead | Fort Lauderdale | 79914 | 1090 |
| S.457.CC.F | 457 | Control Sand | Loggerhead | Fort Lauderdale | 115285 | 1501 |
| S.463.CC.F | 463 | Control Sand | Loggerhead | Fort Lauderdale | 102250 | 1318 |
| S.463.CC.H | 463 | Control Sand | Loggerhead | Hillsboro | 112455 | 1644 |
| S.519.CC.F | 519 | Control Sand | Loggerhead | Fort Lauderdale | 61949 | 955 |
| S.520.CC.F | 520 | Control Sand | Loggerhead | Fort Lauderdale | 74880 | 115 |
| S.521.CC.F | 521 | Control Sand | Loggerhead | Fort Lauderdale | 81443 | 219 |
| S.522.CC.F | 522 | Control Sand | Loggerhead | Fort Lauderdale | 51353 | 916 |
| S.695.CM.H | 695 | Control Sand | Green Turtle | Hillsboro | 100271 | 227 |
| S.696.CM.H | 696 | Control Sand | Green Turtle | Hillsboro | 112269 | 476 |
| S.697.CC.F | 697 | Control Sand | Loggerhead | Fort Lauderdale | 117049 | 220 |
| S.698.CC.F | 698 | Control Sand | Loggerhead | Fort Lauderdale | 132441 | 267 |
| S.699.CC.F | 699 | Control Sand | Loggerhead | Fort Lauderdale | 82399 | 282 |
| S.823.CM.F | 823 | Control Sand | Green Turtle | Fort Lauderdale | 144798 | 192 |
| S.867.CM.F | 867 | Control Sand | Green Turtle | Fort Lauderdale | 139840 | 359 |
| S.892.CC.F | 892 | Control Sand | Loggerhead | Fort Lauderdale | 141714 | 351 |
| U1.1166.CM.H | 1166 | Unhatched Egg | Green Turtle | Hillsboro | 94444 | 121 |
| U1.1167.CM.H | 1167 | Unhatched Egg | Green Turtle | Hillsboro | 120812 | 144 |
| U1.1168.CM.H | 1168 | Unhatched Egg | Green Turtle | Hillsboro | 133301 | 122 |
| U1.127.CC.H | 127 | Unhatched Egg | Loggerhead | Hillsboro | 97504 | 47 |
| U1.236.CC.H | 236 | Unhatched Egg | Loggerhead | Hillsboro | 209641 | 115 |
| U1.242.CC.H | 242 | Unhatched Egg | Loggerhead | Hillsboro | 94836 | 45 |
| U1.282.CC.H | 282 | Unhatched Egg | Loggerhead | Hillsboro | 144418 | 89 |
| U1.322.CC.H | 322 | Unhatched Egg | Loggerhead | Hillsboro | 188918 | 75 |
| U1.323.CC.H | 323 | Unhatched Egg | Loggerhead | Hillsboro | 90928 | 101 |
| U1.392.CC.H | 392 | Unhatched Egg | Loggerhead | Hillsboro | 185816 | 108 |
| U1.456.CC.F | 456 | Unhatched Egg | Loggerhead | Fort Lauderdale | 201276 | 153 |
| U1.457.CC.F | 457 | Unhatched Egg | Loggerhead | Fort Lauderdale | 127385 | 151 |
| U1.463.CC.F | 463 | Unhatched Egg | Loggerhead | Fort Lauderdale | 98948 | 70 |
| U1.463.CC.H | 463 | Unhatched Egg | Loggerhead | Hillsboro | 149631 | 95 |
| U1.519.CC.F | 519 | Unhatched Egg | Loggerhead | Fort Lauderdale | 92312 | 28 |
| U1.520.CC.F | 520 | Unhatched Egg | Loggerhead | Fort Lauderdale | 80073 | 74 |
| U1.521.CC.F | 521 | Unhatched Egg | Loggerhead | Fort Lauderdale | 109555 | 98 |

WILEY

| Samula ID | Nest | Comple true | Turtle eneries | Baach | QIIME sequence | ASV |
|--------------|--------|---------------|----------------|-----------------|-------------------|-----|
| | number | Sample type | | Beach | 29140 | 102 |
| U1 495 CM H | JZZ | | Croop Turtlo | | 129100 | 105 |
| 01.893.CM.H | 695 | | Green Turtle | Hillsboro | 02595 | 170 |
| | 607 | | laggarhaad | Filisboro | 112022 | 75 |
| U1.097.CC.F | 097 | Unhatched Egg | Loggernead | Fort Lauderdale | 113833 | 80 |
| 01.698.CC.F | 698 | | Loggernead | Fort Lauderdale | 128555 | 63 |
| U1.699.CC.F | 699 | | Loggerhead | Fort Lauderdale | 120485 | 92 |
| 01.823.CM.F | 823 | Unhatched Egg | Green Turtle | Fort Lauderdale | 128856 | 129 |
| U1.867.CM.F | 867 | Unhatched Egg | Green Turtle | Fort Lauderdale | 141268 | 93 |
| U1.892.CC.F | 892 | Unhatched Egg | Loggerhead | Fort Lauderdale | 109590 | 94 |
| U2.1166.CM.H | 1166 | Unhatched Egg | Green Turtle | Hillsboro | 129048 | 83 |
| U2.1167.CM.H | 1167 | Unhatched Egg | Green Turtle | Hillsboro | 131661 | 166 |
| U2.1168.CM.H | 1168 | Unhatched Egg | Green Turtle | Hillsboro | 151524 | 169 |
| U2.127.CC.H | 127 | Unhatched Egg | Loggerhead | Hillsboro | 82645 | 52 |
| U2.236.CC.H | 236 | Unhatched Egg | Loggerhead | Hillsboro | 120834 | 99 |
| U2.242.CC.H | 242 | Unhatched Egg | Loggerhead | Hillsboro | 73454 | 112 |
| U2.282.CC.H | 282 | Unhatched Egg | Loggerhead | Hillsboro | 96028 | 98 |
| U2.322.CC.H | 322 | Unhatched Egg | Loggerhead | Hillsboro | 107793 | 102 |
| U2.323.CC.H | 323 | Unhatched Egg | Loggerhead | Hillsboro | 79696 | 179 |
| U2.392.CC.H | 392 | Unhatched Egg | Loggerhead | Hillsboro | 228013 | 105 |
| U2.456.CC.F | 456 | Unhatched Egg | Loggerhead | Fort Lauderdale | 116602 | 102 |
| U2.457.CC.F | 457 | Unhatched Egg | Loggerhead | Fort Lauderdale | 164898 | 249 |
| U2.463.CC.F | 463 | Unhatched Egg | Loggerhead | Fort Lauderdale | 76160 | 206 |
| U2.463.CC.H | 463 | Unhatched Egg | Loggerhead | Hillsboro | 86925 | 87 |
| U2.519.CC.F | 519 | Unhatched Egg | Loggerhead | Fort Lauderdale | 99981 | 83 |
| U2.520.CC.F | 520 | Unhatched Egg | Loggerhead | Fort Lauderdale | 63424 | 97 |
| U2.521.CC.F | 521 | Unhatched Egg | Loggerhead | Fort Lauderdale | 197289 | 102 |
| U2.522.CC.F | 522 | Unhatched Egg | Loggerhead | Fort Lauderdale | 60187 | 101 |
| U2.695.CM.H | 695 | Unhatched Egg | Green Turtle | Hillsboro | 112345 | 82 |
| U2.696.CM.H | 696 | Unhatched Egg | Green Turtle | Hillsboro | 137847 | 107 |
| U2.697.CC.F | 697 | Unhatched Egg | Loggerhead | Fort Lauderdale | 99928 | 101 |
| U2.698.CC.F | 698 | Unhatched Egg | Loggerhead | Fort Lauderdale | 121995 | 98 |
| U2.699.CC.F | 699 | Unhatched Egg | Loggerhead | Fort Lauderdale | 90596 | 74 |
| U2.823.CM.F | 823 | Unhatched Egg | Green Turtle | Fort Lauderdale | 115456 | 120 |
| U2.867.CM.F | 867 | Unhatched Egg | Green Turtle | Fort Lauderdale | 148779 | 99 |
| U2.892.CC.F | 892 | Unhatched Egg | Loggerhead | Fort Lauderdale | 136076 | 65 |
| U3.1166.CM.H | 1166 | Unhatched Egg | Green Turtle | Hillsboro | 107034 | 48 |
| U3.1167.CM.H | 1167 | Unhatched Egg | Green Turtle | Hillsboro | 139796 | 91 |

(Continues)

WILEY_MicrobiologyOpen

TABLE A1 (Continued)

| Sample ID | Nest | Sample type | Turtle species | Beach | QIIME sequence | ASV |
|--------------|------|---------------|----------------|-----------------|-------------------|-----|
| U3 1168 CM H | 1168 | | Green Turtle | Hillsboro | 88387 | 124 |
| U3 127 СС Н | 127 | | Loggerhead | Hillsboro | 61215 | 69 |
| U2 224 CC H | 224 | | Loggerhead | Hillsboro | 04702 | 100 |
| U2 242 CC 11 | 230 | | Loggerneau | | 155027 | 170 |
| U3.242.CC.H | 242 | | Loggernead | Hilisboro | 155937 | 60 |
| U3.282.CC.H | 282 | Unhatched Egg | Loggerhead | Hillsboro | 106699 | 109 |
| U3.322.CC.H | 322 | Unhatched Egg | Loggerhead | Hillsboro | 113420 | 94 |
| U3.323.CC.H | 323 | Unhatched Egg | Loggerhead | Hillsboro | 119792 | 46 |
| U3.392.CC.H | 392 | Unhatched Egg | Loggerhead | Hillsboro | 237925 | 112 |
| U3.456.CC.F | 456 | Unhatched Egg | Loggerhead | Fort Lauderdale | 210742 | 83 |
| U3.457.CC.F | 457 | Unhatched Egg | Loggerhead | Fort Lauderdale | 130341 | 136 |
| U3.463.CC.F | 463 | Unhatched Egg | Loggerhead | Fort Lauderdale | 126119 | 118 |
| U3.463.CC.H | 463 | Unhatched Egg | Loggerhead | Hillsboro | 129293 | 73 |
| U3.519.CC.F | 519 | Unhatched Egg | Loggerhead | Fort Lauderdale | 100580 | 100 |
| U3.520.CC.F | 520 | Unhatched Egg | Loggerhead | Fort Lauderdale | 89481 | 103 |
| U3.521.CC.F | 521 | Unhatched Egg | Loggerhead | Fort Lauderdale | 117780 | 51 |
| U3.522.CC.F | 522 | Unhatched Egg | Loggerhead | Fort Lauderdale | 103251 | 140 |
| U3.695.CM.H | 695 | Unhatched Egg | Green Turtle | Hillsboro | 104053 | 116 |
| U3.696.CM.H | 696 | Unhatched Egg | Green Turtle | Hillsboro | 128937 | 231 |
| U3.697.CC.F | 697 | Unhatched Egg | Loggerhead | Fort Lauderdale | 96902 | 114 |
| U3.698.CC.F | 698 | Unhatched Egg | Loggerhead | Fort Lauderdale | 120225 | 76 |
| U3.699.CC.F | 699 | Unhatched Egg | Loggerhead | Fort Lauderdale | 107036 | 73 |
| U3.823.CM.F | 823 | Unhatched Egg | Green Turtle | Fort Lauderdale | 148461 | 117 |
| U3.867.CM.F | 867 | Unhatched Egg | Green Turtle | Fort Lauderdale | 135124 | 192 |
| U3.892.CC.F | 892 | Unhatched Egg | Loggerhead | Fort Lauderdale | 143452 | 75 |

Note: The number of sequence reads per sample were quantified by QIIME-2 (QIIME Sequence Reads) after sequencing on the Illumina MiSeq. Bacterial taxonomy counts (ASV Count) per sample were determined after the data contamination cleaning process.

<u>MicrobiologyOpen</u>

WILEY

TABLE A2 Alpha diversity pairwise significance results between sample types.

| | Margalef's s Loggerhead | species richness Green Turtle | Fort Lauderdale | Hillsboro |
|-----------------------------|----------------------------|----------------------------------|-----------------|-----------|
| Cloaca, control sand | <0.001 | 0.291 | <0.001 | <0.001 |
| Cloaca, hatched egg | <0.001 | 0.997 | 0.089 | <0.001 |
| Cloaca, nest sand | 0.030 | 0.044 | 0.358 | 0.089 |
| Cloaca, unhatched egg | <0.001 | 1.000 | 0.007 | <0.001 |
| Control sand, hatched egg | <0.001 | 0.247 | <0.001 | <0.001 |
| Control sand, nest sand | <0.001 | 0.924 | <0.001 | <0.001 |
| Control sand, unhatched egg | <0.001 | 0.173 | <0.001 | <0.001 |
| Hatched egg, nest sand | <0.001 | 0.020 | <0.001 | <0.001 |
| Hatched egg, unhatched egg | 0.136 | 0.999 | 0.747 | 0.166 |
| Nest sand, unhatched egg | <0.001 | 0.011 | <0.001 | <0.001 |
| | Shannon div | versity | | |
| | Loggerhead | Green Turtle | Fort Lauderdale | Hillsboro |
| Cloaca, control sand | <0.001 | 0.999 | 0.002 | <0.001 |
| Cloaca, hatched egg | 0.004 | 0.952 | 0.393 | 0.004 |
| Cloaca, nest sand | 1.000 | 0.501 | 0.997 | 0.997 |
| Cloaca, unhatched egg | <0.001 | 0.996 | <0.001 | <0.001 |
| Control sand, hatched egg | <0.001 | 0.843 | <0.001 | <0.001 |
| Control sand, nest sand | <0.001 | 0.347 | 0.006 | <0.001 |
| Control sand, unhatched egg | <0.001 | 1.000 | <0.001 | <0.001 |
| Hatched egg, nest sand | 0.004 | 0.732 | 0.185 | 0.019 |
| Hatched egg, unhatched egg | <0.001 | 0.535 | <0.001 | <0.001 |
| Nest sand, unhatched egg | <0.001 | 0.138 | <0.001 | <0.001 |
| | Inverse Simps | son's diversity | | |
| | Loggerhead | Green Turtle | Fort Lauderdale | Hillsboro |
| Cloaca, control sand | - | - | 1.000 | 0.973 |
| Cloaca, hatched egg | - | - | 0.998 | 1.000 |
| Cloaca, nest sand | - | - | 0.995 | 0.999 |
| Cloaca, unhatched egg | - | - | 0.237 | 0.001 |
| Control sand, hatched egg | - | - | 1.000 | 0.893 |
| Control sand, nest sand | - | - | 0.999 | 0.997 |
| Control sand, unhatched egg | - | - | 0.179 | <0.001 |
| Hatched egg, nest sand | - | - | 1.000 | 0.989 |
| Hatched egg, unhatched egg | - | - | 0.008 | <0.001 |
| Nest sand unbatched egg | _ | _ | 0.081 | <0.001 |

Note: Pairwise tests between the sample types were completed if the "best-fit" GLMM retained the interactive variables (Sample.Type × Turtle.Species, Sample.Type × Beach). If an interaction was not found in the "best-fit" model, no significant values were placed in the table (–). Pairwise tests using the "emmeans" package between the sample types were completed for each turtle species (loggerhead and green turtles) and at each beach (Fort Lauderdale and Hillsboro) for loggerhead samples only. Only significance values (*p* values) are shown in the table. Sample type comparisons that were not found to be significant are indicated in red.

| | TABLE A3 | Alpha diversity pairwise signi | icance results between turtle | e species and beaches f | or each sample type. |
|--|----------|--------------------------------|-------------------------------|-------------------------|----------------------|
|--|----------|--------------------------------|-------------------------------|-------------------------|----------------------|

| | Margalef's species rich | ness | | | | |
|----------------------------|---------------------------|--------------|-------------|-----------|---------------|--|
| | Cloaca | Control sand | Hatched egg | Nest sand | Unhatched egg | |
| Loggerhead, Green Turtle | 0.002 | <0.001 | 0.560 | 0.056 | 0.633 | |
| Fort Lauderdale, Hillsboro | 0.018 | <0.001 | 0.326 | 0.001 | 0.8697 | |
| | Shannon diversity | | | | | |
| | Cloaca | Control sand | Hatched egg | Nest sand | Unhatched egg | |
| Loggerhead, Green Turtle | 0.013 | <0.001 | 0.827 | 0.536 | <0.001 | |
| Fort Lauderdale, Hillsboro | 0.067 | <0.001 | 0.763 | 0.2749 | 0.126 | |
| | Inverse Simpson diversity | | | | | |
| | Cloaca | Control sand | Hatched egg | Nest sand | Unhatched egg | |
| Loggerhead, Green Turtle | - | - | - | - | - | |
| Fort Lauderdale, Hillsboro | 0.895 | 0.563 | 0.707 | 0.966 | 0.131 | |

Note: Pairwise tests between turtle species (loggerhead and green turtles) and beaches (Fort Lauderdale and Hillsboro) for loggerhead samples only were completed for each sample type if the "best-fit" GLMM retained the interactive variables (Sample.Type × Turtle.Species, Sample.Type × Beach). If an interaction was not found in the "best-fit" model, no significant values were placed in the table (–). Only significance values (*p* values) are shown in the table. Sample type comparisons that were not found to be significant are indicated in red.

TABLE A4 One-way ANOSIM pairwise comparisons for all sample types.

| | Lassenhard | | Cus en Tuntle | | Fort Lauderdale | - I- A | Hillsboro | - h- A |
|-----------------------------|-------------|---------|---------------|---------|-----------------|---------|-------------|---------|
| Pairwise comparisons | R Statistic | p value | R Statistic | p value | R Statistic | p value | R Statistic | p value |
| Cloaca, hatched egg | 0.784 | 0.001 | 0.687 | 0.001 | 0.618 | 0.001 | 0.989 | 0.001 |
| Cloaca, nest sand | 0.737 | 0.001 | 0.564 | 0.004 | 0.604 | 0.001 | 0.909 | 0.001 |
| Cloaca, control sand | 0.774 | 0.001 | 0.572 | 0.002 | 0.592 | 0.001 | 0.905 | 0.001 |
| Cloaca, unhatched egg | 0.719 | 0.001 | 0.266 | 0.034 | 0.49 | 0.001 | 0.98 | 0.001 |
| Hatched egg, nest sand | 0.329 | 0.001 | 0.318 | 0.026 | 0.287 | 0.008 | 0.292 | 0.002 |
| Hatched egg, control sand | 0.665 | 0.001 | -0.026 | 0.519 | 0.377 | 0.001 | 0.997 | 0.001 |
| Hatched egg, unhatched egg | 0.345 | 0.001 | 0.132 | 0.001 | 0.262 | 0.001 | 0.421 | 0.001 |
| Nest sand, control sand | 0.496 | 0.001 | 0.275 | 0.003 | 0.235 | 0.002 | 0.985 | 0.001 |
| Nest sand, unhatched egg | 0.123 | 0.031 | -0.079 | 0.693 | 0.072 | 0.196 | 0.132 | 0.096 |
| Control sand, unhatched egg | 0.609 | 0.001 | -0.171 | 0.921 | 0.347 | 0.007 | 0.894 | 0.001 |

Note: ANOSIM R statistics comparing the mean of ranked dissimilarities between groups to the mean of ranked dissimilarities within groups were determined for all possible pairwise comparisons between sample types (cloaca, control sand, hatched egg, unhatched egg, and nest sand). ANOSIM analyses were completed separately by turtle species (loggerhead and green turtle) and beaches (Fort Lauderdale and Hillsboro) for loggerhead samples only. R statistics closer to "1.0" suggest dissimilarity between groups. Pairwise comparisons that were not found to be significant (*p* value) are indicated in red.

VILEY

| | Dissimilarity (%) | Relative abundance (Hatched egg Unhat | %) ched egg | Top contributing bacteria Phylum Contribution (% | ıl phyla) | p value |
|-----------------------------|-------------------|---|----------------|---|---------------|---------|
| Loggerhead | 31.52 | 52.97 | 65.05 | Pseudomonadota | 11.68 | <0.001 |
| | | 26.95 | 14.81 | Bacteroidota | 8.45 | <0.001 |
| | | 15.87 | 15.96 | Bacillota | 8.17 | <0.001 |
| | | 2.41 | 3.72 | Actinomycetota | 2.20 | <0.001 |
| Green Turtle | 18.17 | 66.3 | 65.31 | Pseudomonadota | 7.01 | <0.001 |
| | | 11.78 | 19.75 | Bacillota | 6.27 | 0.002 |
| | | 19.48 | 12.99 | Bacteroidota | 3.60 | <0.001 |
| Fort Lauderdale | 29.01 | 55.78 | 66.19 | Pseudomonadota | 11.10 | 0.001 |
| (Loggerhead only) | | 26.2 | 14.33 | Bacteroidota | 8.08 | <0.001 |
| | | 15.44 | 16.36 | Bacillota | 7.99 | 0.002 |
| | | 1.77 | 2.47 | Actinomycetota | 1.21 | 0.011 |
| Hillsboro (Loggerhead only) | 34.20 | 49.87 | 63.48 | Pseudomonadota | 12.00 | <0.001 |
| | | 27.77 | 15.48 | Bacteroidota | 8.83 | 0.003 |
| | | 16.34 | 15.41 | Bacillota | 8.34 | 0.003 |
| | | 3.11 | 5.44 | Actinomycetota | 3.58 | 0.002 |

TABLE A5SIMPER analyses for hatched and unhatched egg samples at the phylum-level.

Note: SIMPER analyses between hatched and unhatched eggs were completed for each turtle species (loggerhead and green turtles) and at each beach (Fort Lauderdale and Hillsboro) for loggerhead samples only. ASVs were consolidated by phylum before analysis. The total percentage of dissimilarity between egg sample types, relative abundance per egg sample type, and top contributing bacterial phyla (with the percentage of contribution) were determined. Only the bacterial phyla with greater than 1% contribution were selected. The top bacterial phyla contributed a total of 30.50% to the total dissimilarity between hatched and unhatched eggs in loggerhead samples, 16.88% in green turtle samples, 28.38% in Fort Lauderdale samples, and 32.75% in Hillsboro samples.

| | Dissimilarity (%) | Relative abundance Hatched egg Unha | (%) tched egg | Top contributing bacteria Genus Contribution (%) | al genera | p value |
|--------------------------------------|-------------------|--|------------------|---|-----------|---------|
| Loggerhead | 77.00 | 1.31 | 24.80 | Pseudomonas | 11.85 | 0.001 |
| | | 9.84 | 3.90 | Paenibacillus | 4.55 | 0.001 |
| | | 10.95 | 4.58 | Sphingobacterium | 4.47 | 0.001 |
| | | 4.46 | 7.95 | Alcaligenes | 4.28 | 0.012 |
| | | 8.12 | 2.59 | Nitratireductor | 3.68 | 0.001 |
| Green Turtle | 43.68 | 0.72 | 11.23 | Pseudomonas | 5.31 | 0.001 |
| | | 14.65 | 10.43 | Nitratireductor | 3.18 | 0.004 |
| | | 14.32 | 9.73 | Sphingobacterium | 3.10 | 0.006 |
| Fort Lauderdale (Loggerhead only) | 72.39 | 1.40 | 25.49 | Pseudomonas | 12.18 | 0.001 |
| | | 13.29 | 5.12 | Sphingobacterium | 4.91 | 0.001 |
| | | 10.06 | 5.24 | Paenibacillus | 4.75 | 0.001 |
| | | 9.66 | 3.51 | Nitratireductor | 4.08 | 0.002 |
| | | 3.45 | 7.11 | Alcaligenes | 3.47 | 0.039 |
| Hillsboro (Loggerhead only) | 80.86 | 1.21 | 23.85 | Pseudomonas | 11.38 | 0.001 |
| | | 5.57 | 9.11 | Alcaligenes | 5.17 | 0.116 |

TABLE A6 SIMPER analyses for hatched and unhatched egg samples at the genus-level.

(Continues)

TABLE A6 (Continued)

| Dissimilarity (%) | Relative abundance (%) Hatched egg Unhatched egg | | Top contributing bacterial genera Genus Contribution (%) | | p value |
|-------------------|---|------|---|------|---------|
| | 9.61 | 2.07 | Paenibacillus | 4.29 | 0.001 |
| | 8.39 | 3.83 | Sphingobacterium | 3.83 | 0.035 |
| | 6.01 | 2.56 | Flavobacterium | 3.21 | 0.234 |
| | 6.43 | 1.33 | Nitratireductor | 3.06 | 0.006 |

Note: SIMPER analyses between hatched and unhatched eggs were completed for each turtle species (loggerhead and green turtles) and at each beach (Fort Lauderdale and Hillsboro) for loggerhead samples only. The total percentage of dissimilarity between egg sample types, relative abundance per egg sample type, and top contributing bacterial genera (with the percentage of contribution) were determined. Only the bacterial taxa with greater than 3% contribution were selected. Contributing bacterial taxa that were not found to be significant are indicated in red. The top genera contributed a total of 28.83% to the total dissimilarity between hatched and unhatched eggs in loggerhead samples, 11.59% in green turtle samples, 29.39% in Fort Lauderdale samples, and 30.94% in Hillsboro samples.

| TABLE A7 | Relative abundances | of bacteria | identified in | CAP analys | ses for h | natched and | unhatched e | eggs. |
|----------|-----------------------|-------------|----------------|------------|-----------|-------------|-------------|-------|
| | Relative abarraditees | or bucceria | nachten neu ni | Charlany. | 000 101 1 | aterica ana | annacence | -99 |

| | Hatched egg | | Unhatched egg | |
|---------------------|-------------|---------------|---------------|---------------|
| Таха | RA (%) | Variance (±%) | RA (%) | Variance (±%) |
| Phylum | | | | |
| Bacteroidota | 24.83 | 1.543 | 14.29 | 1.353 |
| Verrucomicrobiota | 0.27 | 0.003 | 0.03 | <0.001 |
| Myxococcota | 0.53 | 0.005 | 0.09 | <0.001 |
| Bdellovibrionota | 0.36 | 0.004 | 0.11 | <0.001 |
| Pseudomonadota | 55.82 | 1.810 | 65.00 | 11.769 |
| Bacillota | 14.69 | 0.355 | 16.97 | 2.283 |
| Class | | | | |
| Alphaproteobacteria | 32.47 | 1.019 | 16.04 | 0.671 |
| Bacilli | 13.72 | 0.331 | 9.69 | 1.645 |
| Bacteroidia | 24.82 | 1.543 | 14.29 | 1.353 |
| Gammaproteobacteria | 23.35 | 0.791 | 48.96 | 11.098 |
| Clostridia | 0.89 | 0.023 | 7.08 | 0.635 |
| Order | | | | |
| Sphingomonadales | 3.09 | 0.055 | 0.77 | 0.004 |
| Xanthomonadales | 12.18 | 0.406 | 5.85 | 0.153 |
| Sphingobacteriales | 12.29 | 0.667 | 5.90 | 0.343 |
| Rhizobiales | 24.90 | 0.756 | 13.09 | 0.566 |
| Pseudomonadales | 1.25 | 0.011 | 21.01 | 7.658 |
| Family | | | | |
| Sphingomonadaceae | 3.09 | 0.055 | 0.77 | 0.004 |
| Xanthomonadaceae | 12.07 | 0.405 | 5.82 | 0.153 |
| Sphingobacteriaceae | 12.28 | 0.667 | 5.90 | 0.343 |
| Devosiaceae | 1.77 | 0.014 | 0.76 | 0.007 |
| Pseudomonadaceae | 1.12 | 0.010 | 20.70 | 7.641 |
| Genus | | | | |
| Sphingopyxis | 3.00 | 0.055 | 0.77 | 0.004 |

MicrobiologyOpen _W/ILEV____

31 of 34

TABLE A7 (Continued)

| | Hatched egg | | Unhatched egg | | |
|----------------|-------------|---------------|---------------|---------------|--|
| Таха | RA (%) | Variance (±%) | RA (%) | Variance (±%) | |
| Pseudonocardia | 0.51 | 0.002 | 0.15 | <0.001 | |
| Devosia | 0.85 | 0.006 | 0.21 | 0.001 | |
| Cohnella | 0.24 | <0.001 | 0.06 | <0.001 | |
| Pseudomonas | 1.10 | 0.010 | 19.29 | 7.337 | |

Note: Canonical analysis of principal coordinates (CAP) using Bray-Curtis dissimilarity matrices was used to determine which bacteria best categorize the hatched egg sample types from the unhatched egg sample types. CAP analyses were completed at various taxonomic rankings, from Phylum (top) to Genus (bottom). The relative abundance (RA [%]) of the bacteria identified and their variances (Variance [±%]) were compared between hatched (left two columns) and unhatched (right two columns) eggs. Instances where unhatched eggs had higher abundances are indicated by a red font. Within each taxonomic group, bacteria are organized by the strength of the CAP correlation in categorizing the egg sample types (greater strength at the top of the list).

TABLE A8 SIMPER analysis of sand microbiota between beaches.

| Top contributing bacterial taxa | Hillsboro abundance | Fort Lauderdale abundance | Contribution (%) | p value |
|------------------------------------|------------------------|------------------------------|------------------|---------|
| Sphingobacterium sp. | 0.000 | 0.075 | 4.41 | 0.053 |
| Ochrobactrum sp. | 0.000 | 0.063 | 3.69 | 0.053 |
| Nitratireductor sp. | 0.000 | 0.055 | 3.22 | 0.041 |
| Nitratireductor sp. | 0.000 | 0.044 | 2.58 | 0.053 |
| Achromobacter sp. | 0.001 | 0.033 | 1.92 | 0.052 |
| Brevibacillus sp. | 0.000 | 0.027 | 1.61 | 0.053 |
| Stenotrophomonas sp. | 0.000 | 0.027 | 1.56 | 0.047 |
| Alcaligenes sp. | 0.005 | 0.023 | 1.26 | 0.043 |
| Stenotrophomonas sp. | 0.001 | 0.020 | 1.14 | 0.043 |

Note: SIMPER analyses between Fort Lauderdale and Hillsboro beach were completed for control sand sample types for loggerhead samples only. There was an 84.17% dissimilarity in microbiota between Fort Lauderdale and Hillsboro beach. Only the bacterial taxa with greater than 1% contribution were selected. Abundances of top contributing bacterial taxa at Hillsboro and Fort Lauderdale are listed before the contribution (%) of each bacterial taxon. Contributing bacterial taxa that were not found to be significant are indicated in red.

TABLE A9 Hatched and unhatched egg best-fit generalized linear mixed-effect model.

| | Margalef's species richness | Shannon diversity | Inverse Simpson's diversity |
|---------------------------------|--|---|--|
| Best-fit model variables (dAIC) | Incubation.Length (4.7) High.Tide.Distance (4.0) Temperature.Bottom (8.2) Sample.Type × Turtle.Species (97.4) Sample.Type × Beach (64.1) | Clutch.Size (8.88) High.Tide.Distance (2.59) Temperature.Bottom (4.60) Sample.Type × Turtle.Species (76.37) Sample.Type × Beach (27.81) | Sorting.Coefficient (2.00) Sample.Type × Beach (4.98) |
| R^2 fixed effects alone | 0.76 | 0.64 | 0.25 |
| Model intercept | -61.24 | 0.37 | 0.96 |
| Intercept [95% CI]; t-value | [-154.86, 32.39]; <i>t</i> = -1.29 | [-2.44, 3.17]; <i>t</i> = 0.26 | [0.80, 1.12]; <i>t</i> = 11.95 |

Note: Generalized linear mixed-effect models (GLMMs) were run to determine what metadata variables best explained variation in hatched and unhatched egg samples for each alpha diversity metric (Margalef's species richness, Shannon diversity, and Inverse Simpson's diversity). The best-fit model variables were determined using term selection by AIC score and incorporated a random effect (Nest Number) to account for pseudo-replication. The score difference between the full model and the model without the listed variable (dAIC) is also listed. An R² value for the fixed effects alone is listed to show the explanatory power of the model. The intercept of the model along with its 95% confidence intervals (CI) and *t*-value are also listed. All GLMMs had different variables affecting their variance with varying explanatory power.

UFY_MicrobiologyOpen _

| | Loggerhead | Green Turtle | Fort Lauderdale (Loggerhead only) | Hillsboro (Loggerhead only) |
|--|---|---|---------------------------------------|--|
| BIO-ENV correlation | 0.228 | 0.211 | 0.363 | 0.269 |
| Environmental variables | Clutch size Hatch success Chamber depth High tide distance Conductivity (Bottom) Sand grain size | Incubation length Clutch size Temperature (Bottom) pH (Bottom) | High tide distance Sand grain size | Longitude Hatch success High tide distance Dune distance Temperature (Side) Conductivity (Side) |
| Individual environmental variable (correlation) | Sand grain size (0.152) | Incubation length (0.167) | Sand grain size (0.286) | Conductivity (Side) (0.209) |
| p value | 0.001 | 0.018 | 0.001 | 0.001 |

TABLE A10 Hatched and unhatched egg BIO-ENV correlations for each turtle species and beach.

Note: BIO-ENV correlations were determined for hatched and unhatched eggs for each turtle species (loggerhead and green turtles) and each beach (Fort Lauderdale and Hillsboro) for loggerhead samples only. The highest correlation (BIO-ENV correlation) and the set of environmental variables that go with it are given in the first two rows. The individual environmental variable with the highest correlation is also listed with its correlation in parentheses. All BIO-ENV analyses found significant correlations (p < 0.05) between the identified environmental variables and patterns in bacterial community composition.

ILEY



FIGURE A1 Sample-sized-based rarefaction (interpolation) and prediction (extrapolation) curves by sample type for each turtle species. Turtle species were analyzed separately: loggerhead (top plots) and green turtles (bottom plots). Plots on the left (q = 0) show bacterial taxa richness while plots on the right (q = 2) show Simpson diversity. Diversity was plotted against the number of ASVs ("individuals," sample size) based on sample types: cloaca (red), control sand (brown), hatched egg (green), nest sand (blue), and unhatched egg (purple). The solid portion of the slopes represents the rarefaction (interpolation) of each sample type, while the dashed portion represents the predicted values (extrapolation) of each sample type. A higher diversity value at which the slope reaches a plateau indicates a sample type having greater richness or diversity. Loggerhead samples had greater bacterial richness and diversity than green turtle samples.

2

0 4000 60· Species diversity 3000 40 2000 20 1000 0 0e+00 2e+06 4e+06 6e+06 0e+00 2e+06 4e+06 6e+06 Number of individuals **Hillsboro Samples** 6000 400 300 4000 Species diversity 200 2000 100 0 0 4e+06 0e+00 2e+06 6e+06 0e+00 2e+06 4e+06 6e+06 Number of individuals Cloaca 📥 Control Sand Hatched Egg 💳 Nest Sand Unhatched Egg interpolated extrapolated

Fort Lauderdale Samples

FIGURE A2 Sample-sized-based rarefaction (interpolation) and prediction (extrapolation) curves by sample type for each beach. Beaches were analyzed separately for loggerhead samples only: Fort Lauderdale (top plots) and Hillsboro (bottom plots). Plots on the left (q = 0) show bacterial taxa richness while plots on the right (q = 2) show Simpson diversity. Diversity was plotted against the number of ASVs ("individuals," sample size) based on sample types: cloaca (red), control sand (brown), hatched egg (green), nest sand (blue), and unhatched egg (purple). The solid portion of the slopes represents the rarefaction (interpolation) of each sample type, while the dashed portion represents the predicted values (extrapolation) of each sample type. A higher diversity value at which the slope reaches a plateau indicates a sample type having greater richness or diversity. Hillsboro samples, except for unhatched eggs, had greater bacterial richness and diversity than Fort Lauderdale samples.